Contrasting management techniques and soil types affect network connections between soil properties and the tulip microbiome

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2 Abstract

Conventional agriculture relies on high inputs of chemical fertilisers and pesticides, leading 3 to extensive soil damage. As a result, there has been a shift towards organic cultivation to 4 ameliorate these negative effects. However, knowledge gaps remain concerning how different 5 agricultural management regimes can alter soil properties, root-associated biota and 6 relationships between them. We sampled soils and bulbs from clay and sand tulip fields with 7 8 conventional, organic and hybrid (i.e., intermediate between conventional and organic) management regimes to determine effects on abiotic and biotic soil properties, tulip bulb 9 10 microbiomes/biota and network connectivity between them. We found that soil abiotic properties were not driven by management. However, soil microbial community composition 11 was determined by interactions between soil type and management. Specifically, 12 conventional management created more heterogeneous and homogenous bacterial and fungal 13 soil communities, respectively, across soil types, while organic and hybrid soil communities 14 differed between soil types. A partial least squares path model revealed that across all fields, 15 soil properties determined the composition of tulip bulb biota (i.e., microorganisms, 16 nematodes), while management techniques did not play a significant role, neither directly nor 17 indirectly. Network analyses revealed more modular (i.e., independent) nodes of association 18 in fields with sandy soils. Specifically, management techniques (i.e., chemical usage, 19 ploughing depth, diminished crop rotation) formed a cluster that was independent from most 20 soil (abiotic and biotic) and bulb biotic properties, while clay field networks showed the 21 opposite. Our results indicate that conventional agriculture can homogenise soil microbial 22 communities, with potential impacts on soil function and buffering capacity to stress. 23 Increased modularity between soil properties, management techniques and tulip bulb biota in 24 sand fields could mean increased resistance to disturbance and abiotic and biotic stress as a 25 result of higher functional redundancy. Interactions between agricultural management regime 26

- and soil type must be taken into consideration when determining long-term influences on
- crop microbiome/biota associations and the potential effects on soil health and productivity.

- 30
- 31 Keywords: abiotic-biotic interactions; microorganisms; network analyses; organic
- 32 cultivation; path model; plant microbiome

33 1 | INTRODUCTION

Modern agriculture relies on high inputs of chemical fertilisers, pesticides, nematicides and 34 fungicides necessary to maintain crop health and soil productivity (Tilman et al., 2002). 35 Intensive agriculture increases nitrogen (N) pollution (Richter and Roelcke, 2000), reduces 36 soil biodiversity (Tsiafouli et al., 2015) and destroys soil fertility (Bünemann et al., 2018). As 37 a result, governments are phasing out damaging chemical inputs (Donley, 2019) and 38 encouraging a shift towards organic agricultural (Palm et al., 2014). The EU defines organic 39 agriculture as: "an agricultural method that aims to produce food using natural substances and 40 processes, [thereby promoting] the maintenance of biodiversity, preservation of regional 41 ecological balances and enhancement of soil fertility". These techniques can be used in both 42 food and ornamental crop production. Shifting to organic cultivation typically lowers 43 productivity (Palm et al., 2014), but neutral to positive effects have also been observed 44 (Badgley et al., 2007). However, organic practices improve soil quality (Gomiero et al., 45 2011), enhance ecosystem services (Palm et al., 2014) and provide healthier food (Mie et al., 46 2017). Organic cultivation can affect the soil microbial community, generating higher fungal 47 abundance and network connectivity (Banerjee et al., 2019), with knock-on effects for plant 48 health (Blundell et al., 2020). Nonetheless, many knowledge gaps remain pertaining to the 49 impacts of such methods on the soil, particularly in relation to flower bulb crops (Sukkel and 50 Hommes, 2009). 51

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Healthy soil is the foundation of sustainable, long-term agricultural productivity. Soil
microorganisms such as fungi and bacteria and soil animals such as nematodes, collembola
and mites play an integral role in determining soil characteristics such as structure, food web
composition, nutrient availability and disease level (Ruf and Beck, 2005; Huhta, 2007; Ferris,
2010). Soil communities with a high level of interdependence between taxonomic groups

(i.e., network nodes), but with independence between said nodes (i.e., modularity), create 58 soils that are more resistant to invasion by pathogens (Wei et al., 2015; de Vries and 59 Wallenstein, 2017) and environmental disturbance (de Vries et al., 2012), leading to healthier 60 plants (Bender et al., 2016). Further, management regimes can strongly influence soil 61 network connectivity, with chemical and organic amendments leading to loosely and tightly 62 connected networks, respectively (Ling et al., 2016; Banerjee et al., 2019). The current push 63 by the Dutch and EU governments for more sustainable agricultural practices relies on 64 promoting healthier soils that naturally suppress pests and pathogens (Janvier et al., 2007) 65 and increase the inherent resistance of crops to abiotic and biotic stresses. With integrated 66 knowledge about the relationships between the soil community, soil abiotic properties and 67 plant health, we can build soils that are sustainable and productive. However, current 68 knowledge is lacking about what conditions and management regimes are necessary to create 69 connected, resistant soil networks. 70

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Currently, there is a lack of information on the role soil networks play in determining crop 72 health, in particular that of tulips (Hines et al., 2015). Although a great deal of knowledge 73 exists about the pathogens and pests of tulips (McGovern and Elmer, 2017; Looman et al., 74 2018), the tulip bulb microbiome and biota (e.g., nematodes) that live on and in the bulbs, 75 remains virtually uncharacterised. The very limited knowledge we do have comes from 76 mixed plant communities in the tulip native range (Pershina et al., 2016), meaning it is 77 impossible to draw species-specific conclusions. Further, key information is missing on how 78 management regimes that affect abiotic and biotic soil properties shape the tulip microbiome 79 and the soil network. Acquiring a better understanding of soil networks in tulip fields is the 80 first critical step towards developing more sustainable cultivation systems in the flower bulb 81 industry, with broader application to other crops. 82

Here, we analysed soil abiotic and biotic conditions and tulip microbiomes/biota, across both 84 sand and clay fields that employed conventional, organic and hybrid management regimes. 85 We hypothesised: 1) Contrasting management regimes will generate differences in abiotic 86 and biotic soil properties and bulb biota. This is because organic agriculture relies on reduced 87 chemical pesticide and fertilizer inputs, while generally increasing, e.g., the addition of 88 organic matter and the frequency of crop rotation schemes and these management techniques 89 are strong determinants of soil properties (Gomiero et al., 2011) and the microbiomes (Arif et 90 91 al., 2020). Further, management techniques (e.g., chemical use, crop rotation) are expected to have direct effects on tulip bulb biota and indirect effects mediated through soil abiotic and 92 biotic properties (Figure A.1); 2) Fields with contrasting soil texture (i.e., sand versus clay) 93 will have different responses in abiotic and biotic soil properties to management regime (i.e., 94 contrasting interactions between management regime and soil type). This is because clay 95 soils, as compared to sand soils, have different physical and structural conditions such as poor 96 percolation, reduced oxygen content (Boekel, 1970) and abiotic characteristics (Dexter, 97 2004), leading to effects on soil-bound pests and diseases (McGovern and Elmer, 2017; 98 Looman et al., 2018) and plant-associated biota (Fierer, 2017); and 3) Across sand versus 99 clay soils, management techniques will create different levels of connectivity between soil 100 abiotic and biotic properties and tulip bulb biota due to inherent differences caused by soil 101 texture (Boekel, 1970; Dexter, 2004) and the known impact of agricultural management on 102 soil networks (de Vries and Wallenstein, 2017). 103

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105 2 | MATERIALS AND METHODS

106 2.1 | Experimental setup

Eighteen tulip field representing different soils types and management regimes were selected 107 from North Holland and Flevoland in The Netherlands (Table A.1). Farms were selected that 108 use organic, hybrid and conventional management. Organic farms followed EU regulations 109 for organic cultivation (Meredith and Willer, 2014). Hybrid farms used fewer artificial 110 fertilizers, pesticides, herbicides, etc., but were not certified organic. Conventional farms 111 used standardized agricultural methods, which included chemical inputs. For each 112 management type, farms were selected that have sandy soils ($\geq 75\%$ sand) and clay (>10%) 113 clay) soils. In total, this resulted in: 3 management types \times 2 soils types \times 3 replicates = 18 114 115 fields.

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117 2.2 | Soil and bulb sampling

Between 22 May and 4 June 2020, soils and bulbs were collected from 20 locations ranging 118 from approximately 15-30 meters apart across each field in a "W" shaped pattern to capture 119 sufficient heterogeneity (Sabbe and Marx, 1987). Soil samples used to measure all abiotic 120 and biotic soil characteristics (except mites and collembola; see below) were collected from 121 each location to a depth of 25 cm using a 2.4 cm diameter corer. The top 25 cm of soil is 122 considered the most biologically active layer of agricultural soils (Fan et al., 2016a) and 123 encompasses the majority of the rooting depth of tulip bulbs (de Hertogh et al., 1983). Five 124 bulbs were excavated from each of the 20 sampling locations with a hand trowel. Soil was 125 gently shaken off, the leaves and stems removed and bulbs were placed into a paper sack. At 126 each location, a penetrometer measurement was taken to determine soil compaction 127 (Sanglerat, 2012). Samples used for extraction of mites and collembola were collected with a 128 4 cm diameter corer to a depth of 5 cm because the majority of mite and collembola are found 129 in this upper layer of soil (Römbke et al., 2006). These samples were collected from ten 130 points across each field. 131

133 **2.3 | Farmer interviews**

134 Farmers were interviewed regarding management techniques such as crop rotation, pesticide

use, number of years cultivated, etc. (Table S2).

136

137 2.4 | Soil abiotic properties

138 Soil abiotic (and a few of the biotic) property analyses were performed by Eurofins

139 (Wageningen, The Netherlands) using standard soil abiotic characteristic procedures. For

140 details, see Appendix A.1.

141

142 **2.5 | Chemical residue analyses**

143 A subsample of soil and bulb material from each field was analysed for chemical residues.

144 Samples were analysed by Eurofins (Wageningen, The Netherlands) using the Mini Luke

145 method (Lozano *et al.*, 2016).

146

147 **2.6 | Nematode extractions**

148 Nematodes were extracted from soils using a modified version of the Baermann funnel

149 method and from bulbs using a mist chamber (Van Bezooijen, 2006). The number of

nematodes in each sample was determined using an inversion microscope (×40-200

151 magnification). The first 100 nematodes encountered in the counting dish were separated into

152 different feeding groups: plant-feeding, bacterial-feeding, fungal-feeding and omnivore-

153 carnivore (Yeates *et al.*, 1993). For details, see Appendix S2.

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155 2.7 | Mite, collembola and Diptera extraction

156	Soil samples used to characterise mites, collembola and Diptera were kept refriger	rated (5 °C)
157	for one to three days before processing. For details, see Appendix S2.	
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159	2.8 Virus detection in the tulip bulbs
160	The tulip bulbs were tested in the Bloembollenkeuringsdienst (BKD) laboratory (Lisse, The
161	Netherlands) for the presence of Arabis mosaic virus (ArMV), cucumber mosaic virus
162	(CMV), lily symptomless virus (LSV), tulip breaking virus (TBV), tulip virus X (TVX),
163	tobacco rattle virus (TRV), tobacco necrosis virus (TNV) and tulip mosaic virus (TuMV).
164	The tests were performed using the double-antibody sandwich enzyme-linked
165	immunosorbent assay (DAS-ELISA) (Luo et al., 2012) using in-house produced antisera.
166	
167	2.9 DNA extraction and sequencing
168	The microbial community of soil and bulb samples was analysed using high-throughput
169	sequencing. The extraction and isolation of total DNA and the PCRs were done at WUR
170	Glastuinbouw laboratory in Bleiswijk, The Netherlands. Analyses were performed by Utrecht
171	Sequencing Facility (USEQ, UMCU, Utrecht, The Netherlands). For details, see Appendix
172	S3.
173	
174	2.10 Statistical analyses
175	2.10.1 ANOVAs
176	Each response variable (chemical, physical, biological, farmer interview responses) was
177	analysed using ANOVA with management regime (organic, hybrid, conventional) and soil type
178	(clay, sand) as fixed factors. All data were checked to ensure all assumptions for ANOVA were
179	met and transformed as necessary (see ANOVA Tables for details). When significant effects
180	were detected, data were subjected to post-hoc tests (Day and Quinn, 1989) using the

emmeans/multcomp packages in R (Lenth, 2019) with Tukey HSD adjustment for multiple
comparisons. The ANOVAs were performed using R (R Core Team, 2020) with the packages
lme4/lmerTest (Bates *et al.*, 2015; Kuznetsova *et al.*, 2017).

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185 2.10.2 | Sequencing data

To process the raw sequences, for both the ITS and 16S reads, Cutadapt (Martin, 2011) was used to remove forward and reverse primers and allow the separation of the fungal reads from the bacterial reads. All raw sequencing data were de-noised with DADA2 (Callahan *et al.*, 2016) following the DADA2 Pipeline Tutorial (1.18). For details, see Appendix S4.

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191 2.10.3 | Partial least squares path model

We used a partial least squares path model (PLS-PM) to investigate the direct and indirect relationships between management regimes, soil abiotic and biotic properties, and tulip biota (Sanchez *et al.*, 2013). This type of structural equation modelling (SEM) is becoming more prominent in natural science (Fan *et al.*, 2016b) and has no distributional or independence requirements (Fornell and Bookstein, 1982; Chin, 2010). For details, see Appendix S5.

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198 2.10.4 | Network and heatmap analyses

The correlation networks and heatmaps were made in R 3.6 (R Core Team, 2020). Networks
were based on Veen *et al.* (2019) and heatmaps were made using the cor() and heatmap.2()
functions of the R-package "gplots" (Warnes *et al.*, 2020). For details, see Appendix S5.

202

203 **3 | RESULTS**

204 3.1 | Soil abiotic properties

Soil abiotic properties from clay and sand tulip fields showed many differences (Tables S3S4). Values from the penetrometer, total N, N supplying capacity, total and available S, total
K, Ca and Mg, pH, organic C, soil organic matter, carbonic lime, clay, silt, clay humus and
Ca saturation were higher on clay soils. Values for sand, Mg, K and Na saturation and soil
friability and slaking were higher on sand soils. No abiotic soil properties differed
significantly between management regimes nor were there any significant interactions
between soil type and management.

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213 **3.2** | Soil and bulb biotic properties

Several soil biotic properties differed by soil type and/or management. The tulip virus tests
revealed extremely low infection rates, making these data unfit for statistical analyses. For
details, see Appendix S6. See appendices for details on effects on the nematode (Appendix
S6, Tables S5-S8) and microarthropod communities (Appendix S6, Tables S9-A.10, Figure
S2).

219

Soil and bulb microbial communities differed between soil type and/or management regime. 220 Here, we present the results germane to our hypotheses, with further details in Appendix S6, 221 Figures S3-S8, Tables A.11-A.12, Supporting Excel Document. Bulb bacterial and fungal 222 community composition were not affected by soil type or management (Figures 1a, 2a). The 223 soil bacterial community was affected by soil type, with communities from clay soils forming 224 a distinct cluster (Figure 1b, Table A.11). However, bacterial communities from organic and 225 hybrid clay fields clustered together, while organic sand field communities formed another 226 distinct cluster and conventionally managed fields showed wide heterogeneity regardless of 227 soil type. Soil fungal communities showed a slightly different pattern: communities from 228 organic and hybrid clay fields formed a cluster separate from their counterparts on sand soils, 229

while fungal communities from conventionally managed clay and sand fields formed a
distinct, relatively homogenous cluster (Figure 2b, Table A.11). A number of bacterial and
fungal diversity metrics were affected by soil type and management, but no significant
interactions were observed (Appendix S6, Table A.12, Figures S3-S4).

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The Deseq2 analyses revealed differences between soil and bulb microbial taxa driven by soil 235 type and management (Figures S5-S8, Supporting Excel Document). Soil type did not 236 significantly affect taxonomic abundances of bulb bacterial communities. Bacillus and 237 Streptomyces were higher in sand soils, while Fusarium and Purpureocillium were higher in 238 clay soils (Figure S5). Overall, 20 and 16 fungal taxa were more abundant in organic versus 239 conventional and hybrid bulbs, respectively (Figure S6a,b). *Trichoderma* was higher in bulbs 240 from organic versus conventional and hybrid (Figure S6a,b). Botrytis and Fusarium were 241 lower in bulbs from conventional versus hybrid fields (Figure S6c) and Fusarium was more 242 abundant in organic versus conventional and hybrid soils (Figure S8a,b). 243

244

245 **3.3 | Management regime properties**

Crop rotation was more frequent on clay versus sand soils and on organic versus hybrid and 246 conventional fields. However, sand fields with hybrid and conventional management had 247 more infrequent crop rotation versus organic sand fields and all management regimes on clay 248 fields (soil type × management interaction) (Tables A.13-A.14, Figure S9a). A selection of 249 answers to the interview questions is shown in Table A.1. Of particular noteworthiness, 5 of 250 6, 1 of 6 and 1 of 6 of the organic, hybrid and conventional farmers added animal manure, 251 respectively. Chemical use and residue in soils and bulbs was higher in conventional and 252 hybrid versus organic (none) fields. For details, see Tables A.15-A.16, Figures A.10-A.11. 253 254

3.4 | Partial least squares path model

The final PLS-PM showed abiotic and biotic soil properties explained significant variation in 256 tulip bulb biota (Figure 3, $R^2 = 0.66$). Soil abiotic properties were directly negatively driven 257 by management techniques (NB: considered as individual measures, not categorically) (-258 0.47). Management directly negatively influenced soil biotic properties (-0.63), but also 259 exerted a weak, indirect negative effect mediated through soil abiotic properties (-0.05). Soil 260 abiotic properties did not significantly directly affect soil biotic properties (-0.11). Soil 261 abiotic properties had negative direct (-0.55) and indirect (-0.05) effects on tulip bulb biota 262 mediated through soil biotic properties. Soil biotic properties had a marginally significant (p 263 < 0.10) direct positive effect (0.42) on tulip bulb biota. Finally, management regime did not 264 have a direct effect on tulip bulb biota, but an insignificant negative effect mediated via soil 265 abiotic and biotic properties (-0.38). For loadings and original and bootstrapped path 266 coefficients, see Tables A.17-A.18. 267

268

269 **3.5** | Network analyses and heatmaps

The network analysis across different soils types showed connections between soil abiotic 270 and biotic properties, management regime techniques and tulip bulb biota differed between 271 sand and clay soils (Figure 4, Table A.19). Here, we discuss some of the more remarkable 272 patterns. Overall, networks on clay soils were more compact, with multiple links between 273 numerous variables and few distinct nodes, while sand soils showed a looser network, with 274 fewer links between variables and more distinct, separate nodes. In clay networks, the bulb 275 bacterial and fungal taxa were present in nodes that were distinct from nodes that contained 276 bulb chemical resides, but in sand networks, bulb microorganisms were present in nodes 277 more distant from one another, and some were even associated with bulb chemical residues. 278 Further, bulb-associated nematodes in clay soils were loosely clustered, while in sand soils 279

they clustered tightly. Management techniques (i.e., chemical residues, ploughing depth, crop
rotation) formed a distinct node in sandy soils, but in clay soils, these parameters broke apart
into several different nodes. The six panels to the right in Figure 4 show pair-wise
comparisons of how the values of the measured parameters changed between management
regimes in clay and sandy soils. Many of the changes between management regime were
similar across sand versus clay fields. For correlation heatmaps, see Figures A.12-A.13.

287 4 | DISCUSSION

We found that abiotic soil properties were driven by soil type and not management. However, 288 the relative abundances of certain microbial bulb and soil taxa were affected by soil type and 289 management. The soil microbial community composition was determined by an interaction 290 between soil type and management regime. A PLS-PM showed a marginally significant 291 indirect, but no direct, effects of management techniques (i.e., contributions of specific 292 techniques, not categorically divided into management regime) on tulip bulb biota, but strong 293 effects of management techniques on soil abiotic and biotic properties. Network analysis 294 showed strong differences in connectivity between soil abiotic and biotic and tulip bulb biotic 295 properties and management techniques across soil types. Below we discuss what these results 296 mean in the context of agricultural management regimes. 297

298

299 4.1 | Effects of management on soil properties and tulip bulb biota

In contrast to our first hypothesis, we found no differences in abiotic soil properties between
organic, hybrid and conventional management regimes. Many studies found differences in
abiotic soil properties emerge rapidly under organic management (Clark *et al.*, 1998;
Gomiero *et al.*, 2011), which contrasts our findings. However, correlation analyses revealed
that fewer years between cycles of tulip cultivation resulted in depletion of virtually all soil

nutrients. Of the six fields with the lowest frequency in rotation between tulip crops, tulips
were grown the previous year, which may have depleted total soil N (Abasi *et al.*, 2016). On
two of the remaining four fields with tight crop rotation, narcissus and lily were grown the
previous year. Lilies have a high demand for N (Slangen *et al.*, 1989) and narcissus have a
high demand for K, Mg and P (Hanks, 2002) relative to tulips (Niedziela *et al.*, 2015; Abasi *et al.*, 2016), This suggests that considering management categorically my partially mask the
effects of specific management techniques (i.e., crop rotation).

312

313 In partial support of our first hypothesis, we found an effect of management on bulb and soil microorganisms. Fusarium and Botrytis abundances, two important tulip fungal diseases 314 (Looman et al., 2018), were higher in bulbs from hybrid versus conventional fields and 315 Fusarium was more abundant in organic versus conventional and hybrid soils. This could be 316 attributed to two of six hybrid fields having grown tulips the previous year, with these 317 diseases remaining in the soil the following growing season (McGovern and Elmer, 2017) 318 (NB: no differences in Fusarium and Botrytis were found between hybrid versus 319 conventional soils). Lower fungicide application in hybrid and organic versus conventional 320 fields may have caused these differences, highlighting the need to find alternatives to 321 chemicals to combat fungal diseases. In addition, Trichoderma spp. abundance was higher in 322 organic versus hybrid and conventional bulbs. Trichoderma spp. can colonise roots, thereby 323 outcompeting pathogens for space (Harman, 2000) and can enhance tulip flower quality and 324 speed up bulb maturation (Mazhabi et al., 2011). Resultantly, compared to hybrid and 325 conventional, organic tulip bulbs might have higher resistance to disease. Finally, more 326 groups of fungal taxa were more abundant in organic versus hybrid and conventional bulbs. 327 This suggests that organic bulbs host a wider range of fungi, but precisely what this means 328 for bulb health remains unknown. 329

331	The PLS-PM revealed that tulip bulb biota were not significantly driven by direct or indirect
332	effects of management regime (NB: considered as individual <i>techniques</i> and not
333	categorically). This may have occurred for a number of reasons such as tulip bulbs recruiting
334	their own rhizosphere microbiome via root exudates (Berg and Smalla, 2009) or the
335	confounding effect of different cultivars (Jiang et al., 2017). Instead, soil abiotic properties
336	exerted significant direct and indirect effects mediated via soil biotic properties, while soil
337	biotic properties exerted a nearly significant direct effect on tulip bulb biota. It is intuitive
338	that the soil community would be a driver of tulip bulb microbiota since organisms present in
339	bulk soil colonise bulbs and roots (Berg and Smalla, 2009). Future emphasis should be placed
340	on what specifically drives soil and bulb microbiomes towards enhanced resistance and
341	resilience (Berendsen <i>et al.</i> , 2012).

342

343 4.2 | Interactive effects between management and soil type on soil properties and tulip 344 bulb biota

In partial support of our second hypothesis, we found that interactions between soil type and 345 management influenced certain soil biotic properties. Abiotic properties remained unaffected, 346 possibly due to inherent differences between clay and sandy soils overriding interactions with 347 management (Boekel, 1970). Bulb biota were also unaffected, with soil abiotic and biotic 348 properties acting as stronger determinants. Soil nematodes and arthropods were unaffected by 349 management, possibly due in part to seasonal effects on soil animal populations (Werner and 350 Dindal, 1990). However, bacterial communities from organic soils formed two distinct, tight 351 clusters driven by soil type, but showed considerable heterogeneity in conventional soils. 352 Bacterial communities differ markedly by soil type in arable systems (Girvan et al., 2003), 353 and these distinctions were retained in organic clay versus sand soils, but lost in conventional 354

soils. Fungal communities showed patterns similar to bacterial communities in organic soils, 355 but conventional fungal communities formed a relatively homogenous cluster across soils 356 types. These differences may be due to different life strategies. Bacteria typically grow more 357 quickly with shorter generation times (Rousk and Bååth, 2011) and are less resistant to 358 disturbance (Sun et al., 2017). Therefore, under conventional management in both clay and 359 sand fields, bacterial communities may have diverged as they recovered from disturbance 360 (e.g., chemical input, deeper ploughing) based on random founder effects (Kim et al., 2013), 361 while fungal communities ended up homogenised across soils types perhaps due to a 362 combination between certain taxa being eliminated or enhanced by disturbance (Schnoor et 363 al., 2011) and higher resistance to disturbance (Sun et al., 2017). Contrasting responses 364 between bacterial and fungal communities may have implications for soil functioning (Berga 365 et al., 2012), tulip health and productivity. 366

367

368 4.3 | Soil network connectivity across management regimes and soil types

In support of our third hypothesis, network connectivity between soil properties, management 369 techniques and bulb biota differed between soil types. Overall, networks on clay soils were 370 more compact with fewer distinct nodes compared to those on sand soils. Compact networks 371 can benefit soil ecosystem services, such as carbon storage (Morriën et al., 2017). Clay soils 372 typically contain more organic matter than sand soils, indicating clay fields have greater 373 potential to sequester carbon, thereby providing benefits for climate change-related goals in 374 agriculture (Zaehle et al., 2007). In contrast, there is evidence that soil networks broken into 375 distinct nodes (i.e., modules) have a greater buffering capacity to stress (e.g., drought, pests, 376 pathogens) due to functional redundancy amongst microorganisms (Wei et al., 2015; de Vries 377 and Wallenstein, 2017), which aligns with the higher modularity between distinction nodes in 378 sand soil networks. This supports existing evidence that disease levels are lower in tulip bulbs 379

grown in sand soils (McGovern and Elmer, 2017; Looman et al., 2018) and was also shown 380 here with greater Fusarium abundance in clay versus sand soils. Further, management 381 techniques (i.e., chemical input, ploughing, crop rotation) formed a distinct node in sand soils 382 and showed fewer associations with bulb and soil biota. Distinct nodes could contribute to 383 improved tulip bulb harvests on sandy soils because intensive management practices might be 384 "compartmentalised", thereby isolating the negative impacts of intensive management on soil 385 biota. Reversed, in clay soils, intensive management techniques exert influence across 386 multiple soil properties, thereby potentially playing a disruptive role. The next step is to 387 discover how to steer soil networks into the appropriate direction to suppress disease and 388 maximise ecosystem services. 389

390

391 4.4 | Conclusions

We found that management regime did not affect soil abiotic properties, but soil and tulip 392 bulb microbial taxa were driven by management. Bacterial and fungal soil communities 393 formed distinct clusters with organic and hybrid management on clay soils. However, 394 conventional management resulted in heterogeneous and homogenous bacterial and fungal 395 communities, respectively, regardless of soil type. Sandy soil networks showed more distinct 396 nodes, while this was reversed on clay soils. This could increase buffering capacity in sandy 397 soils against abiotic and biotic stresses and disturbances via increased functional redundancy. 398 These findings pull focus on the need to consider interactive effects between soil type and 399 management on the soil microbial community and soil networks. Future research should 400 explore how long-term changes to management regimes interact with different soil types and 401 crop microbiomes/biota and thereby alter the functional relationships between soil network 402 connectivity and crop health and productivity. 403

405	Authors'	contributions
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- 406 J.R.D.L., C.B., G.M. and P.v.L. conceived the experimental design; J.R.D.L., C.B., A.P.,
- 407 P.v.L. and K.P. collected the data; J.R.D.L., E.L.F., L.B.S. and M.D. analysed the data;
- 408 J.R.D.L. led the writing of the manuscript. All authors contributed to revising the manuscript.

410 **Data availability**

411 Data will be made available in the Dryad Digital Data Repository upon acceptance.

412

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416

417 **Conflict of interest**

418 The authors declare no conflict of interest.

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420 **References**

- Abasi, H., Babalar, M., Lessani, H., Naderi, R., 2016. Effects of nitrogen form of nutrient
 solution on uptake and concentration macro element and morphological trait in
- 423 hydroponic tulip. Journal of Plant Nutrition 39, 1745-1751.
- Arif, I., Batool, M., Schenk, P.M., 2020. Plant microbiome engineering: Expected benefits
 for improved crop growth and resilience. Trends in Biotechnology.
- 426 Badgley, C., Moghtader, J., Quintero, E., Zakem, E., Chappell, M.J., Aviles-Vazquez, K.,
- 427 Samulon, A., Perfecto, I., 2007. Organic agriculture and the global food supply.
 428 Renewable Agriculture and Food Systems 22, 86-108.
- Banerjee, S., Walder, F., Büchi, L., Meyer, M., Held, A.Y., Gattinger, A., Keller, T., Charles,
 R., Van Der Heijden, M.G., 2019. Agricultural intensification reduces microbial
- network complexity and the abundance of keystone taxa in roots. The ISME journal 13,
 1722-1736.

- Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects models
 using lme4. 2015 67, 1-48.
- Bender, S.F., Wagg, C., van der Heijden, M.G., 2016. An underground revolution:
- biodiversity and soil ecological engineering for agricultural sustainability. Trends in
 ecology & evolution 31, 440-452.
- Berendsen, R.L., Pieterse, C.M., Bakker, P.A., 2012. The rhizosphere microbiome and plant
 health. Trends in plant science 17, 478-486.
- Berg, G., Smalla, K., 2009. Plant species and soil type cooperatively shape the structure and
 function of microbial communities in the rhizosphere. FEMS microbiology ecology 68,
 1-13.
- Berga, M., Szekely, A.J., Langenheder, S., 2012. Effects of disturbance intensity and
 frequency on bacterial community composition and function. Plos One 7, 11.
- Blundell, R., Schmidt, J.E., Igwe, A., Cheung, A.L., Vannette, R.L., Gaudin, A.C., Casteel,
- 446 C.L., 2020. Organic management promotes natural pest control through altered plant
 447 resistance to insects. Nature plants 6, 483-491.
- Boekel, P., 1970. Soil structure problems in tulip culture. International Symposium on
 Flowerbulbs 23, pp. 338-343.
- Bünemann, E.K., Bongiorno, G., Bai, Z., Creamer, R.E., De Deyn, G., de Goede, R.,
- Fleskens, L., Geissen, V., Kuyper, T.W., Mäder, P., 2018. Soil quality–A critical
 review. Soil Biology and Biochemistry 120, 105-125.
- 453 Callahan, B.J., McMurdie, P.J., Rosen, M.J., Han, A.W., Johnson, A.J.A., Holmes, S.P.,
- 454 2016. DADA2: High-resolution sample inference from Illumina amplicon data. Nature
 455 Methods 13, 581-583.
- Chin, W.W., Dibbern, J., 2010. Handbook of partial least squares concepts, methods and
 applications. Springer, Heilderberg.
- Clark, M.S., Horwath, W.R., Shennan, C., Scow, K.M., 1998. Changes in soil chemical
 properties resulting from organic and low-input farming practices. Agronomy Journal
 90, 662-671.
- Day, R.W., Quinn, G.P., 1989. Comparisons of treatments after an analysis of variance in
 ecology. Ecological Monographs 59, 433-463.
- de Hertogh, A.A., Aung, L.N., Benschop, M., 1983. The tulip : botany, usage, growth and
 development. CABO, Wageningen.

- de Vries, F.T., Liiri, M.E., Bjørnlund, L., Bowker, M.A., Christensen, S., Setälä, H.M.,
- Bardgett, R.D., 2012. Land use alters the resistance and resilience of soil food webs to
 drought. Nature Climate Change 2, 276-280.
- de Vries, F.T., Wallenstein, M.D., 2017. Below-ground connections underlying above-ground
 food production: a framework for optimising ecological connections in the rhizosphere.
 Journal of Ecology 105, 913-920.
- 471 Dexter, A.R., 2004. Soil physical quality: Part I. Theory, effects of soil texture, density, and
 472 organic matter, and effects on root growth. Geoderma 120, 201-214.
- 473 Donley, N., 2019. The USA lags behind other agricultural nations in banning harmful
 474 pesticides. Environmental Health 18, 44.
- Fan, J., McConkey, B., Wang, H., Janzen, H., 2016a. Root distribution by depth for
 temperate agricultural crops. Field Crops Research 189, 68-74.
- Fan, Y., Chen, J., Shirkey, G., John, R., Wu, S.R., Park, H., Shao, C., 2016b. Applications of
 structural equation modeling (SEM) in ecological studies: an updated review.
 Ecological Processes 5, 1-12.
- Ferris, H., 2010. Contribution of nematodes to the structure and function of the soil food web.
 Journal of Nematology 42, 63-67.
- Fierer, N., 2017. Embracing the unknown: disentangling the complexities of the soil
 microbiome. Nature Reviews Microbiology 15, 579.
- Fornell, C., Bookstein, F.L., 1982. Two structural equation models: LISREL and PLS applied
 to consumer exit-voice theory. Journal of Marketing research 19, 440-452.
- 486 Girvan, M.S., Bullimore, J., Pretty, J.N., Osborn, A.M., Ball, A.S., 2003. Soil type is the
- 487 primary determinant of the composition of the total and active bacterial communities in
 488 arable soils. Applied and environmental microbiology 69, 1800-1809.
- 489 Gomiero, T., Pimentel, D., Paoletti, M.G., 2011. Environmental impact of different
- agricultural management practices: conventional vs. organic agriculture. Critical
 reviews in plant sciences 30, 95-124.
- Hanks, G.R., 2002. Commercial production of Narcissus bulbs. Narcissus and daffodil: the
 genus Narcissus, 53-130.
- Harman, G.E., 2000. Myths and Dogmas of Biocontrol Changes in Perceptions Derived from
 Research on Trichoderma harzinum T-22. Plant Disease 84, 377-393.
- 496 Hines, J., van der Putten, W.H., De Deyn, G.B., Wagg, C., Voigt, W., Mulder, C., Weisser,
- 497 W.W., Engel, J., Melian, C., Scheu, S., 2015. Towards an integration of biodiversity–

- 498 ecosystem functioning and food web theory to evaluate relationships between multiple
 499 ecosystem services. Advances in ecological research 53, 161-199.
- Huhta, V., 2007. The role of soil fauna in ecosystems: A historical review. Pedobiologia 50,
 489-495.
- Janvier, C., Villeneuve, F., Alabouvette, C., Edel-Hermann, V., Mateille, T., Steinberg, C.,
 2007. Soil health through soil disease suppression: Which strategy from descriptors to
 indicators? Soil Biology & Biochemistry 39, 1-23.
- 505 Jiang, Y., Li, S., Li, R., Zhang, J., Liu, Y., Lv, L., Zhu, H., Wu, W., Li, W., 2017. Plant
- cultivars imprint the rhizosphere bacterial community composition and association
 networks. Soil Biology and Biochemistry 109, 145-155.
- Kim, M., Heo, E., Kang, H., Adams, J., 2013. Changes in soil bacterial community structure
 with increasing disturbance frequency. Microbial ecology 66, 171-181.
- Kuznetsova, A., Brockhoff, P.B., Christensen, R.H.B., 2017. ImerTest package: tests in linear
 mixed effects models. 2017 82, 26.
- Lenth, R., 2019. emmeans: Estimated Marginal Means, aka Least-Squares Means. R package
 version 1.3.5.1.
- Ling, N., Zhu, C., Xue, C., Chen, H., Duan, Y., Peng, C., Guo, S., Shen, Q., 2016. Insight
 into how organic amendments can shape the soil microbiome in long-term field
- experiments as revealed by network analysis. Soil Biology and Biochemistry 99, 137149.
- Looman, B., van Leeuwen, P., Slootweg, G., van Dam, M., Vreeburg, P., 2018. Ziekten en
 plagen bij Bloembollen.
- Lozano, A., Kiedrowska, B., Scholten, J., de Kroon, M., de Kok, A., Fernández-Alba, A.R.,
 2016. Miniaturisation and optimisation of the Dutch mini-Luke extraction method for
 implementation in the routine multi-residue analysis of pesticides in fruits and
- vegetables. Food chemistry 192, 668-681.
- Luo, Y., Terkawi, M.A., Jia, H., Aboge, G.O., Goo, Y.-K., Cao, S., Li, Y., Yu, L., Ooka, H.,
 Kamyingkird, K., 2012. A double antibody sandwich enzyme-linked immunosorbent
 assay for detection of secreted antigen 1 of Babesia microti using hamster model.
 Experimental parasitology 130, 178-182.
- Martin, M., 2011. Cutadapt removes adapter sequences from high-throughput sequencing
 reads. EMBnet. journal 17, 10-12.

Mazhabi, M., Nemati, H., Rouhani, H., Tehranifar, A., Moghadam, E.M., Kaveh, H., 2011. 530 How may Trichoderma application affect vegetative and qualitative traits in tulip 531 "darwin hybride" cultivar. 532 McGovern, R.J., Elmer, W.H., 2017. Diseases of Tulip. In: McGovern, R.J., Elmer, W.H. 533 (Eds.), Handbook of Florists' Crops Diseases. Springer International Publishing, Cham, 534 pp. 1-26. 535 Meredith, S., Willer, H., 2014. Organic in Europe-prospects and developments. IFOAM EU 536 Group. 537 Mie, A., Andersen, H.R., Gunnarsson, S., Kahl, J., Kesse-Guyot, E., Rembiałkowska, E., 538 Quaglio, G., Grandjean, P., 2017. Human health implications of organic food and 539 organic agriculture: a comprehensive review. Environmental Health 16, 111. 540 Morriën, E., Hannula, S.E., Snoek, L.B., Helmsing, N.R., Zweers, H., de Hollander, M., Soto, 541 R.L., Bouffaud, M.-L., Buée, M., Dimmers, W., Duyts, H., Geisen, S., Girlanda, M., 542 Griffiths, R.I., Jørgensen, H.-B., Jensen, J., Plassart, P., Redecker, D., Schmelz, R.M., 543 Schmidt, O., Thomson, B.C., Tisserant, E., Uroz, S., Winding, A., Bailey, M.J., 544 Bonkowski, M., Faber, J.H., Martin, F., Lemanceau, P., de Boer, W., van Veen, J.A., 545 van der Putten, W.H., 2017. Soil networks become more connected and take up more 546 carbon as nature restoration progresses. Nature Communications 8, 14349. 547 Niedziela, C.E., Nelson, P.V., Dickey, D.A., 2015. Growth, Development, and Mineral 548 Nutrient Accumulation and Distribution in Tulip from Planting through Postanthesis 549 Shoot Senescence. International Journal of Agronomy 2015, 341287. 550 Palm, C., Blanco-Canqui, H., DeClerck, F., Gatere, L., Grace, P., 2014. Conservation 551 agriculture and ecosystem services: An overview. Agriculture Ecosystems & 552 Environment 187, 87-105. 553 Pershina, E., Ivanova, E., Nagieva, A., Zhiengaliev, A., Chirak, E., Andronov, E., Sergaliev, 554 N.K., 2016. A comparative analysis of microbiomes in natural and anthropogenically 555 disturbed soils of northwestern Kazakhstan. Eurasian Soil Science 49, 673-684. 556 R Core Team, 2020. R: A Language and Environment for Statistical Computing. 557 Richter, J., Roelcke, M., 2000. The N-cycle as determined by intensive agriculture - examples 558 from central Europe and China. Nutrient Cycling in Agroecosystems 57, 33-46. 559 Römbke, J., Sousa, J.-P., Schouten, T., Riepert, F., 2006. Monitoring of soil organisms: a set 560 of standardized field methods proposed by ISO. European Journal of Soil Biology 42, 561 S61-S64. 562

- Rousk, J., Bååth, E., 2011. Growth of saprotrophic fungi and bacteria in soil. FEMS
 Microbiology Ecology 78, 17-30.
- Ruf, A., Beck, L., 2005. The use of predatory soil mites in ecological soil classification and
 assessment concepts, with perspectives for oribatid mites. Ecotoxicology and
 Environmental Safety 62, 290-299.
- Sabbe, W.E., Marx, D.B., 1987. Soil sampling: spatial and temporal variability. Soil testing:
 Sampling, correlation, calibration, and interpretation 21, 1-14.
- Sanchez, G., Trinchera, L., Russolillo, G., 2013. plspm: tools for partial least squares path
 modeling (PLS-PM). R package version 0.4 1.
- 572 Sanglerat, G., 2012. The penetrometer and soil exploration. Elsevier.
- Schnoor, T.K., Lekberg, Y., Rosendahl, S., Olsson, P.A., 2011. Mechanical soil disturbance
 as a determinant of arbuscular mycorrhizal fungal communities in semi-natural
 grassland. Mycorrhiza 21, 211-220.
- Slangen, J.H.G., Krook, G.J., Hendriks, C.H.M., Hof, N.A.A., 1989. Nitrogen dressing and
 nutrient absorption of lilies (asiatic hybrids) on sandy soils. Netherlands Journal of
 Agricultural Science 37 (1989); ISSN: 0028-2928. 1989.
- Sukkel, W., Hommes, M., 2009. Research on organic agriculture in the Netherlands:
 organisation, methodology and results. Wageningen UR.
- 581 Sun, S., Li, S., Avera, B.N., Strahm, B.D., Badgley, B.D., 2017. Soil bacterial and fungal
- communities show distinct recovery patterns during forest ecosystem restoration.Applied and Environmental Microbiology 83.
- Tilman, D., Cassman, K.G., Matson, P.A., Naylor, R., Polasky, S., 2002. Agricultural
 sustainability and intensive production practices. Nature 418, 671-677.
- 586 Tsiafouli, M.A., Thébault, E., Sgardelis, S.P., de Ruiter, P.C., van der Putten, W.H.,
- Birkhofer, K., Hemerik, L., de Vries, F.T., Bardgett, R.D., Brady, M.V., Bjornlund, L.,
- Jørgensen, H.B., Christensen, S., D' Hertefeldt, T., Hotes, S., Hol, W.H.G., Frouz, J.,
- Liiri, M., Mortimer, S.R., Setälä, H., Tzanopoulos, J., Uteseny, K., Pižl, V., Stary, J.,
- Wolters, V., Hedlund, K., 2015. Intensive agriculture reduces soil biodiversity across
 Europe. Global Change Biology 21, 973-985.
- Van Bezooijen, J., 2006. Methods and techniques for nematology. Wageningen University
 Wageningen, The Netherlands.
- Veen, G., Snoek, B.L., Bakx-Schotman, T., Wardle, D.A., van der Putten, W.H., 2019.
- Relationships between fungal community composition in decomposing leaf litter and
 home-field advantage effects. Functional Ecology 33, 1524-1535.

Warnes, G.R., Bolker, B., Bonebakker, L., Gentleman, R., Huber, W., Liaw, A., Lumley, T.,
Maechler, M., Magnusson, A., Moeller, S., 2020. gplots: various R programming tools
for plotting data. R Package Version 3.0.4. R Foundation for Statistical Computing
Vienna, Austria.
Wei, Z., Yang, T., Friman, VP., Xu, Y., Shen, Q., Jousset, A., 2015. Trophic network
architecture of root-associated bacterial communities determines pathogen invasion and
plant health. Nature communications 6, 1-9.
Werner, M.R., Dindal, D.L., 1990. Effects of conversion to organic agricultural practices on
soil biota. American Journal of Alternative Agriculture, 24-32.
Yeates, G.W., Bongers, T., Degoede, R.G.M., Freckman, D.W., Georgieva, S.S., 1993.
Feeding-habits in soil nematode families and genera- an outline for soil ecologists.
Journal of Nematology 25, 315-331.
Zaehle, S., Bondeau, A., Carter, T.R., Cramer, W., Erhard, M., Prentice, I.C., Reginster, I.,
Rounsevell, M.D., Sitch, S., Smith, B., 2007. Projected changes in terrestrial carbon
storage in Europe under climate and land-use change, 1990–2100. Ecosystems 10, 380-

612 401.

613 **Figure legends**

FIGURE 1. Bulb (a) and soil (b) bacterial community composition in samples from tulip
fields with organic, hybrid and conventional management on sand and clay soils. For F- and
p-values, see Table A.11.

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FIGURE 2. Bulb (a) and soil (b) fungal community composition in samples from tulip fields
with organic, hybrid and conventional management on sand and clay soils. For F- and pvalues, see Table A.11.

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FIGURE 3. Partial least squares path model showing the relationships between management regime, soil abiotic properties, soil biotic properties and tulip bulb biota in tulip fields with clay and sand soil with organic, hybrid and conventional management regimes. The width of the arrows indicates the strength of the causal relationships supplemented by standardised path coefficients (~p < 0.10; *p < 0.05; **p < 0.01; ***p < 0.001). R^2 values indicate the explained variance of the response variables. For details on measured indicator variables, loadings and path coefficients, see Appendix S5, Tables A.17-A.18.

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FIGURE 4. Network analysis showing the interactions between soil abiotic and biotic
properties and tulip bulb biota in clay and sand fields. Pairwise comparisons (six panels on
the right hand side) indicate higher abundance in fields with the first mentioned management
regime when variables are green and higher abundance in fields with the second mentioned
management regime when variables are magenta, with brighter colours indicating higher
abundances. See Table A.19 for numeric coding of variables.