

Global projections for anthropogenic reactive nitrogen emissions to the atmosphere: an assessment of scenarios in the scientific literature

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Most long-term scenarios of global reactive nitrogen (Nr) emissions to the atmosphere are produced by Integrated Assessment Models in the context of climate change assessments. These scenarios indicate that these global Nr emissions are likely to increase in the next decades, followed by a stabilization or decline. Crucial factors for future Nr emissions are the development of the underlying drivers (especially fertilizer use, animal husbandry, transport, power generation and fires), air pollution control and climate policies. The new scenarios made for climate change research and assessment, the Representative Concentration Pathways – RCPs, cover a smaller range of possible Nr emission projections than the literature, as they all assume progressive air pollution control. A more focused development of scenarios for air pollution may be needed to improve both the relevance and quality of the scenarios for research and assessment of air pollution (and possibly short term climate change).

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Introduction

Disturbance of the nitrogen (N) cycle has been identified as a key sustainability problem in several international environmental assessments [1–4]. The most important reactive nitrogen substances emitted into the atmosphere by human activities are nitrous oxide (N₂O), nitric oxides (NO and NO₂ together denoted as NO_x) and ammonia (NH₃) (we will refer to these together as Nr gases).

Nitrogen input to surface and subsurface waters are an additional form of disturbance of the N-cycle, but this paper will focus on atmospheric emissions. An important consequence of atmospheric emissions is the increased deposition of nitrogen oxides, ammonia and ammonium species, leading to several environmental effects. This includes 1) acidification, 2) excess N loading in ecosystems (eutrophication) leading to decreased biodiversity, 3) direct toxicity, 4) nitrate leaching, and 5) increased susceptibility to secondary stress factors and ecosystem diversity [5,6,7,8,9].

The contribution of increased N deposition to acidification and eutrophication has been analysed at a global scale [7,10] showing that both acidification and eutrophication can be expected to expand outside industrialized countries. Excess Nr loading is expected to lead to subsequent shifts in plant species composition towards nitrophilic species [11–13], which in turn, may lead to biodiversity loss, and in aquatic systems to algal blooms and decreased water quality. Increased Nr deposition may also lead to increased carbon storage [9,14,15] and thus affecting the carbon cycle. Finally, Nr gas emissions can play a role in the greenhouse effect. N₂O emission contribute to about 6% of global greenhouse gas emissions [16], while Nr emissions can also affect tropospheric ozone chemistry and aerosol formation, which both play an important role in climate change forcing [17].

Given the importance of Nr gas species for these environmental effects, it is important to assess how mankind will influence the global N cycle in the future. Many different processes are responsible for Nr gas emissions. Fossil fuel combustion, biomass burning, lightning and microbiological processes in soils are the major processes involved in the production of NO_x [18] and N₂O [19]. NH₃ stems mostly from volatilization from animal waste and synthetic fertilizers, biomass burning, losses from oceans and soils under natural vegetation, emissions from waste, industrial processes and traffic [20]. As discussed further in this article, many of the economic activities that drive Nr emissions are expected to increase over time. At the same time, increasing environmental awareness and improvements in technology could lead to emission reduction. While for greenhouse gas emission scenarios (most importantly CO₂) regularly model comparisons are organised [21–23], similar systematic comparisons for future N emissions do not exist (and also the number

of projections is considerably less). Recently, Granier *et al.* [24] compared different emission inventories and showed considerable uncertainties even with respect our understanding of historic and current emissions.

In this paper, we present a brief assessment based on literature on expected future changes in reactive N gas emissions under different scenario assumptions. We will focus primarily at the global level. It should be noted, however, that many more projections exist at the regional as part of initiatives to improve regional environmental quality. At the global level, many of the scenarios evaluated here were performed in the context of assessment of greenhouse gas emissions (as climate models use reactive N gas emissions as input data). In Section 'Scenario analysis', we first discuss some of the key concepts in the scenario literature. In Section 'Projections of atmospheric nitrogen emissions', we look into the projections for NO_x, N₂O and NH₃. Finally, in Section 'Conclusions' we draw the conclusions of this assessment.

Scenario analysis

Integrated assessment models (IAM) are often used to develop global emission scenarios. Integrated assessment models looking specifically at air pollution often have a regional focus, such as the RAINS for Europe [25,26] and the RAINS Asia model for South and East Asia [27,28]. The RAINS model has more recently been expanded into the GAINS model expand to joint strategies for managing air pollution and greenhouse gas emissions [29,30]. Integrated Assessment models of climate change often focus at the global scale. Their scenarios often also include Nr gas emissions. Some of these IAMs are also applied to study environmental problems in a more general context, such as for the Global Environmental Outlook and the Millennium Ecosystem Assessment [1–3].

Future developments in Nr gas emissions are estimated on the basis of projected changes in relevant economic activities and the emissions per unit of these activities. This can be done in different ways. The most common approach is to calculate emission on the basis of the product of economic activity levels and emission factor (Emission = Activity × emission_factor). The latter equals the emission rate per unit of the activity. Such emission factors can be estimated on the basis of detailed representations of abatement technologies ('technology basis') combined with rules on the desired environmental quality or maximum expenditures [26,31]. Emissions factors can, however, also be single values as a function of time, that follow certain exogenously prescribed or endogenously derived trends representative for the particular sector or region ('representative emission factors'). Both the simple and more complex derived emission factors can be determined on the basis of existing and future policies in different parts of the world (such as the Current Legislation, CLE) scenarios of Cofala *et al.* [30] or the use of empirically

observed trends such as the Environmental Kuznets Curve (EKC) [32–35]. The EKC suggests that, starting from low-income levels, emissions will originally increase with increasing income but at some point will peak and subsequently decline. The latter is driven not only by increasingly tight environmental policies, but also by shifts towards industries with lower emissions (include the service sector) and improved technology. There is, however, a fundamental debate whether the EKC is valid and whether it can be extrapolated to the future [32,36]. Usually emission factors will change over time to reflect changes in technology, policy and economic activity levels. As discussed in Section 'Current emissions uncertainty', there is a wide set of emission inventories that contain information on current emissions factors. For Nr gas emissions from soils where biological processes (nitrification and denitrification in the case of N₂O and NO) or physical-chemical processes (NH₃ volatilization) are responsible for the production, consumption and emission, often more complex approaches are used. Such approaches range from statistical models [37,38] to more complex mechanistic models [39,40].

The uncertainties in the processes that determine future emissions as discussed above (economic activity levels, technological development, future legislation, and its effective implementation, chemical and physical interactions) contribute to the uncertainty in future emissions. In most cases, Nr gas emissions do not occur as a fixed fraction of some input or activity level (as is the case for CO₂), which leads to a larger amount of uncertainty in these emissions. In the case of combustion emissions, Nr gas emissions depend on combustion temperature and other characteristics of the technology considered. Emissions from biological processes are extremely variable in space and time, and depend on temperature, soil properties, and soil moisture.

When comparing scenarios, uncertainty becomes apparent in different ways. First of all, different models and studies lead to different estimates of future emissions. Secondly, even within one study a model may be used to develop widely diverging scenarios based on different assumptions on the type of future scenario [30,41] or the introduction of climate policy [42]. In the discussion in the remainder of this article, the spread in emissions projections (indicated by the 10th–90th percentile of their range) is loosely interpreted as an indication of uncertainty in future emissions levels. It is useful to keep in mind, however, that future emissions may well fall outside the range of scenario projections, particularly since many of the scenarios examined were not constructed with a principal focus on Nr gas emissions. It goes beyond the scope of this brief assessment to discuss the development of individual emission factors in detail. In order to better understand the differences in the projections, it would be useful to do so in a more detailed paper also in relation to the emission factors

used in inventories of current emissions (see also Section ‘Current emissions uncertainty’).

For the purpose of this assessment, we have compared global scenarios that include a description of long-term (i.e. at least several decades) trends in emissions of NO_x , N_2O and NH_3 . Studies included within this category are: 1) some of the scenarios that are developed to study air pollution, such as [30,43], 2) the IPCC-SRES scenarios [44], 3) the set scenarios from a wide range of models used by Van Vuuren *et al.* [45^{*}] to project 21st century climate change (based on the EMF-21 scenarios [22]), 4) the scenarios developed for EMF-22 [21] and 5) the Representative Concentration Pathways recently published to examine the implications of a range of future climate forcings [46,47^{**}]. We also discuss scenarios developed to explicitly assess different futures with respect to global environmental change, such as the scenarios related to the N cycling as developed for the Millennium Ecosystem Assessment [41,48,49^{*},50^{*}] and the International Assessment of Agricultural Knowledge, Science and Technology for Development (IAASTD) [3,50^{*}]. The Integrated Model to Assess the Global Environment (IMAGE) model [51] has been used to develop several sets of such scenarios in different assessments. We acknowledge that a large number of scenarios have been developed for regional assessments of the impact of Nr gas emissions [28,52]. These scenarios, however, fall outside the scope of this assessment.

Projections of atmospheric nitrogen emissions

Current emissions uncertainty

The emission scenarios examined here, are generally based on emission inventories for a specific base year as a starting point. There is range of different inventories that provide emission data on a global level. Important available inventories include the EDGAR database [53], the database underlying the RAINS/GAINS system (see earlier references), the RETRO database [54] and the ACCMIP database. The latter has actually been constructed on the basis of combination and harmonization of published and publicly available datasets [24,55^{*}]. A recent overview of available inventories was conducted by Granier *et al.* [24]. This overview shows that, in general, there are large differences across the databases, on the global scale but even more so on the regional scale (see also [56]). Among the different Nr gas, there are also clear differences in the degree of consensus among historical emission and inventories. This uncertainty in the base-year and historical emissions is a major source of uncertainty in future projections.

The global N_2O emission budget is reasonably well-constrained by available knowledge on the most important sink, N_2O destruction in the atmosphere. Even with this overall constraint, the partitioning between

anthropogenic and natural N_2O emissions is difficult to quantify. There are few such overall observationally based constraints for NO_x or NH_3 emissions. Yet, Granier *et al.* [57] conclude that “there is a rather good consensus on the NO_x global emissions”. For the year 1980, for instance, it was found that the difference between three databases was equal to 13%, while for the year 2000, the five inventories compared show a maximum difference around 15–20%, respectively.

Most of the inventories do not include explicit uncertainty estimates [24]. Schöpp *et al.* [58] estimated a 10–20% uncertainty at the country level for 1990 NO_x emissions in Europe, and slightly higher values for NO_x emissions from shipping in European waters, and NH_3 emissions in general. Beusen *et al.* [59] estimated a 20% uncertainty in NH_3 emissions from agricultural production, where uncertainty in individual NH_3 sources (such as animal manure storage, spreading, grazing animals) may be much larger. The foundation of these uncertainty assessments is expert judgment, since measurement data necessary for a more rigorous approach is generally not available (and real uncertainty might even be larger).

The comparison of existing inventories suggests that the uncertainty might be considerably higher [24]. One factor that reduces uncertainty for emissions such as NO_x is cancellation across sectors. Errors in emissions coefficients for one sector are generally expected to be uncorrelated with errors in another sector, which reduces the net uncertainty. Overall, the uncertainty is expected to be larger in countries without well-developed infrastructure for emissions inventory development (in some cases, there might be factor of 2 differences between observations and model calculations).

A detailed comparison of emission factors in the various inventories and future scenarios is currently. Although this exercise is far from straightforward (e.g. owing to different sectoral definitions), it would be of value in understanding differences between various studies and, potentially, reducing uncertainty

Projections for NO_x

A recent estimate of global anthropogenic NO_x emissions in the year 2000 amounted to nearly 40 Tg N yr^{-1} [55^{*}]. These emissions rapidly increased during the 20th century. The major part of these emissions (more than 30 Tg yr^{-1}) originate from fossil fuel combustion, that is, road transport (10 Tg), shipping and aviation (6 Tg), the energy sector (7.5 Tg), industry (4.5 Tg) and buildings (3 Tg). In addition, around 6 Tg N yr^{-1} originates from biomass burning.

In the future, the most important sources, that is, transport and power production, are expected to grow rapidly in terms of energy consumption, in particular in developing

countries. As a result, it is likely that these sectors will continue to dominate NO_x emissions. Typically, global energy demand by the transport sector is projected to grow by about 60–260% across a range of scenarios in the 2000–2050 period and expected to grow even more in the second half of the century [34,42,60]. For electricity production, scenarios show even faster increases: 200–300% in the 2000–2050 period and possibly at a similar rate after 2050.

While the growth in these activity levels would lead to higher NO_x emissions, other factors could partly offset this such as the increasing stringency of NO_x air pollution control policies in different parts of the world and the fuel and technology choices (including efficiency) in the energy system (which could be influenced by climate policy, see below).

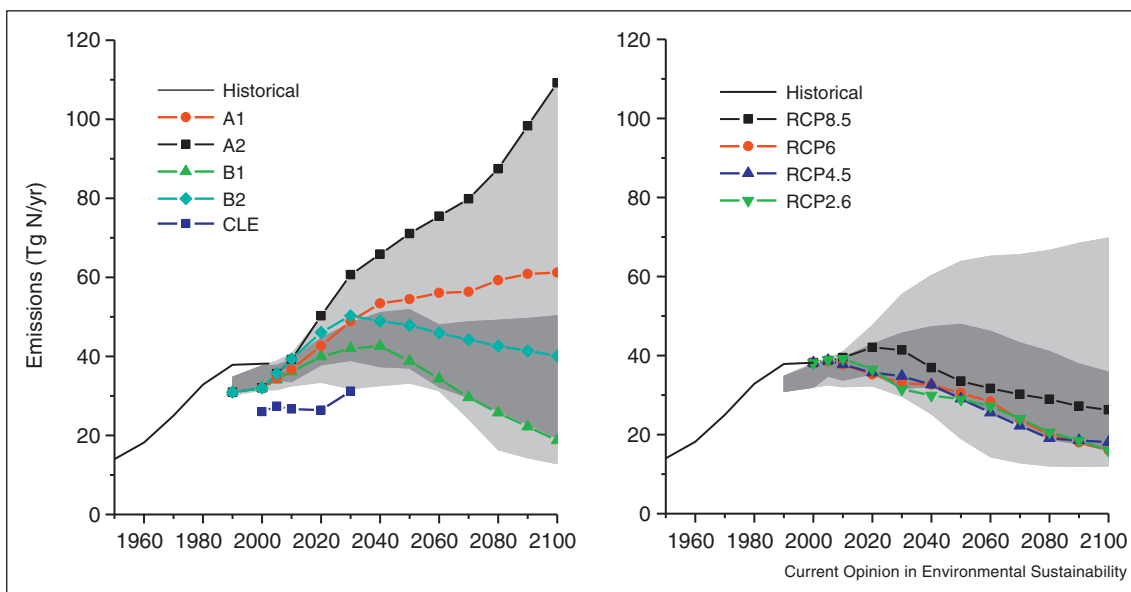
Future emission scenarios span a wide range from about 10 up to 100 Tg N yr^{-1} and higher (Figure 1). According to Cofala *et al.* [30] the current legislation scenario would more-or-less stabilize global emissions in the next decades, while the maximum feasible reductions would result in a 60% reduction compared to 2000. The IPCC-SRES A2 scenario published in 2000 [44] and a more recently published scenario by MIT [61] are at the high end of the range (70 Tg N yr^{-1} in 2050 and above 100 Tg N yr^{-1} in 2100). Most scenarios, however, show emissions to be in the order of 30–50 Tg N yr^{-1} in 2050 and 20–50 Tg N yr^{-1} in 2100 – or in other words, a 10–20% increase in the next few decades followed by a

stabilization of emissions (without climate policy) or a modest decline (with climate policy). This implies that the range is still rather comparable as the range considered some years ago, such as by Unger *et al.* [62] (31–56 Tg N yr^{-1} in 2030).

Emissions are expected to grow much less than the corresponding economic drivers, as discussed above. This is the result of considerable decreases in future emission factors in most scenarios. It can be easily derived that, at the global level, for median scenarios the ‘aggregated emission factors’ (as a result of air pollution control and technology development) decrease by around 60% in 2050 and 80% in 2100. Frozen emission factors, that is, constant at 2000 level, would result in annual emissions in the order of 120 Tg N and 200 Tg N in 2050 and 2100, respectively. As the more detailed calculations underlying the Cofala projections indicate, such reductions are technically possible.

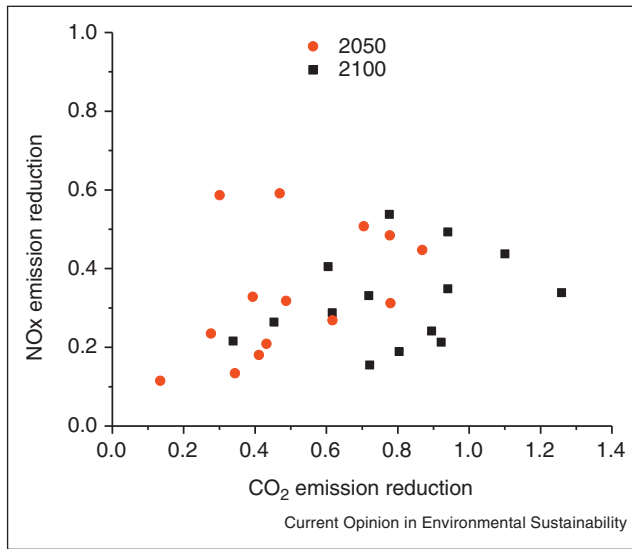
Comparison of the left-hand and right-hand side of Figure 1 shows that also future climate policy is likely to reduce NO_x emissions, given the fact that the systemic changes to reduce CO_2 emissions introduced in the energy system in response to climate policy, will also reduce NO_x emissions. Data from a set of scenarios from 6 different models suggest that on average a 10% reduction in CO_2 emissions leads to a 5% reduction in NO_x emissions (Figure 2). The available data also suggest that this ‘co-benefit’ ratio is somewhat reduced in the long-term (less NO_x reduction for a given reduction of CO_2).

Figure 1



Future NO_x emissions according to various scenarios (light grey area covers the 10–90th percentile; dark grey area the 25–75th percentile). The right hand panel only includes scenarios without climate policy (22 scenarios); the left hand panel includes the full set of scenarios (with and without climate policy) (40 scenarios). The graph also shows the scenarios of the IPCC-SRES set [37], the IIASA-CLE scenario (both sets do not include climate policy) [26] and the RCPs (including climate policy) [40].

Figure 2



Relationship between CO₂ and NO_x emission reduction (fraction of baseline emissions) as a result of the introduction of climate policy in the scenario set used by van Vuuren *et al.* [38], created by 6 different models. The figure indicates that in all models, climate policy alone also leads to significant NO_x emission reduction (although the exact relationship differs per model).

Typically, scenarios from literature portray very clear shifts in emissions from OECD countries to Asia and to a lesser degree other developing regions, as illustrated in Figure 3 for one scenario. Similar trends are observed in more detailed regional projections, such as for Europe and

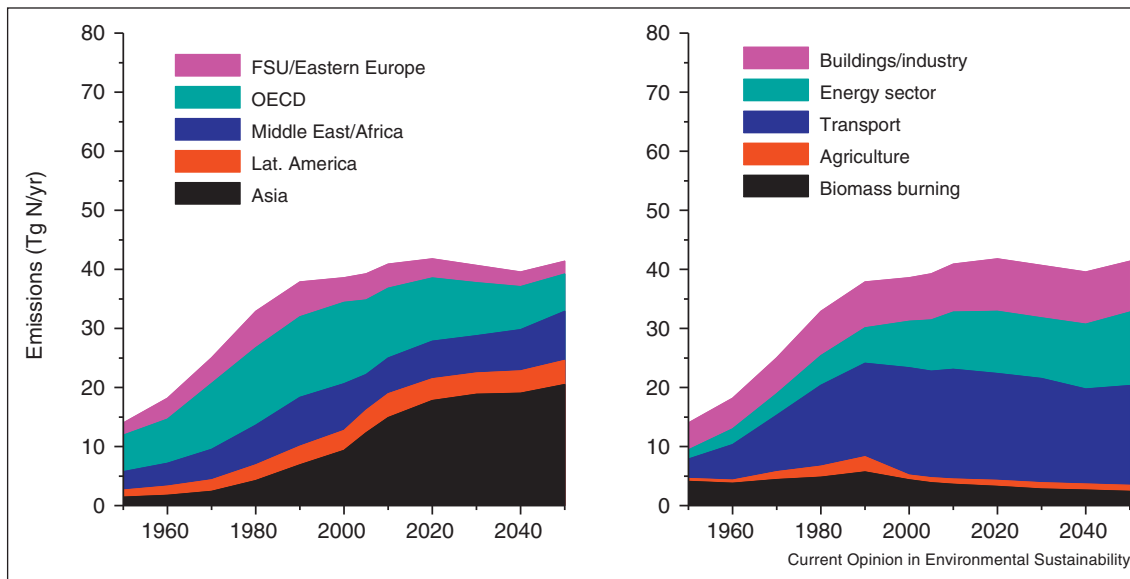
Asia [28,30,63,64]. In terms of sectors, emissions mostly increase in the energy sector (power generation). Transport emissions remain more-or-less stable as a result of opposing trends in activity levels and emission factors, following a fast-global technology shift, imposed by regional emission standards such as EURO2-to-EURO6.

Projections for N₂O

Estimates of anthropogenic N₂O emissions in 2000 are of the order of 7.5 Tg N yr⁻¹ [55]. The bulk of these emissions stem from agricultural activities (7 Tg N yr⁻¹), especially direct and indirect emissions of fertilizer use and animal husbandry. About 0.5–1 Tg N yr⁻¹ originates from the energy and industry sector, that is, the production of nitric and adipic acid and the transport sector.

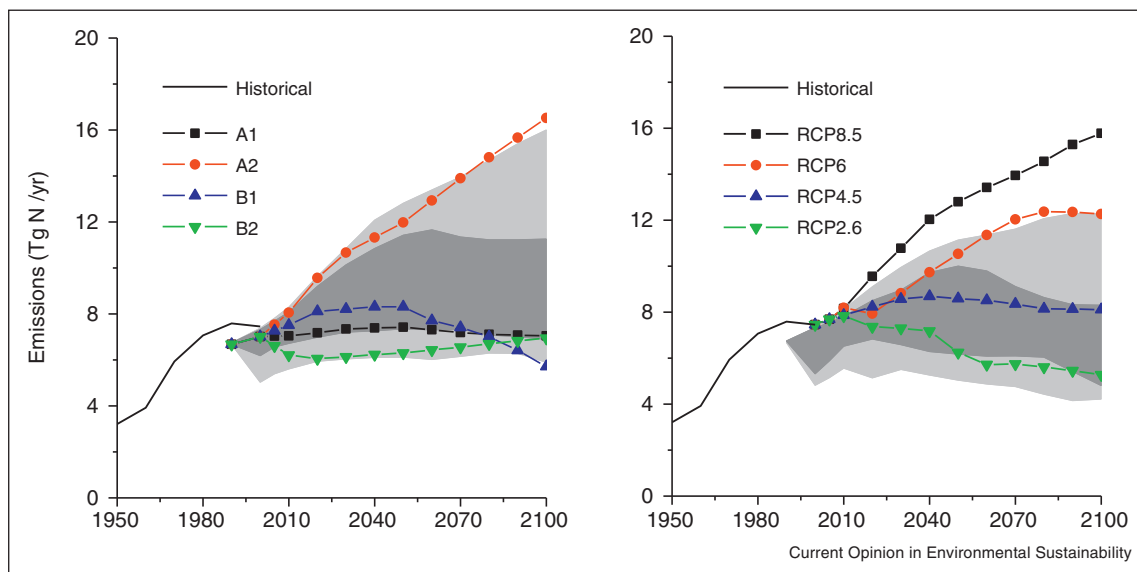
N fertilizer use and animal husbandry are expected to continue to grow slowly in most scenarios. Scenarios by the IMAGE model, for instance, show typically a 50–100% increase in N fertilizer use and a 50–150% increase in livestock production during the 21st century [65–67]. By contrast, N₂O emission from industrial sources are expected to decline significantly, as it is relatively easy to control these emissions in case of climate policy; in fact, important reductions have already been achieved in the past decades in different parts of the world [68]. Emissions from transport are relatively small – and thus has only a small impact on overall N₂O emissions, although they have been slowly increasing with the introduction of catalytic converters. Future N₂O emissions will thus depend on future agricultural production and practices and climate policy. Key agricultural factors include the

Figure 3



NO_x emissions according to the IMAGE reference scenario as published by van Vuuren *et al.* [56].

Figure 4



Future N_2O emissions according to various scenarios (light grey area covers the 10–90th percentile; dark grey area the 25–75th percentile). The right hand panel only includes scenarios without climate policy; the left hand panel includes the full set of scenarios (with and without climate policy). In addition, the graph shows the scenarios of the IPCC-SRES set and the RCPs (including climate policy) (sources see Figure 1).

shifts in dietary patterns (the degree of meat consumption) and fertilizer application [67].

Figure 4 shows the range of scenarios from literature. In the absence of climate policy, most scenarios expect N_2O emissions to increase somewhat (around 0–40%) in the 2000–2050 period followed by stabilization (partly as result of a stabilizing global population). Some scenarios, however, indicate a further rapid increase in N_2O emissions after 2050 (such as the IPCC A2 scenario, that projects a further population increase after 2050 [44]).

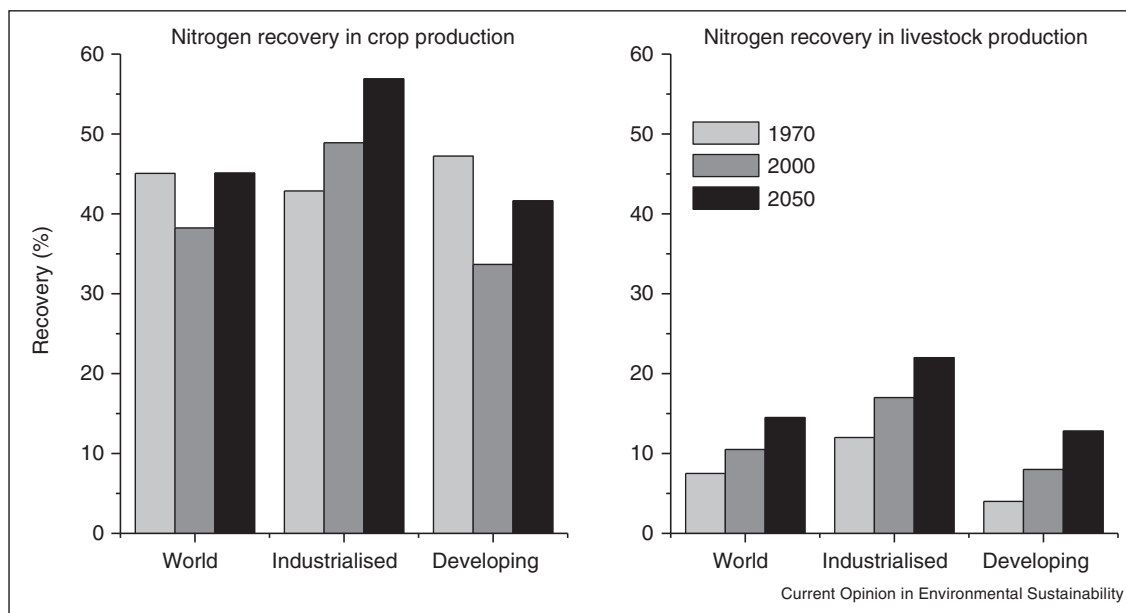
These results indicate that growth rates of N_2O emissions without climate policy are not very different from those of the underlying economic factors while scenarios for NO_x show a reduction of the emission factors owing to other factors than climate policy, this is a less important factor for future N_2O emissions, although some changes will occur owing to changes in underlying practices. Improved knowledge of the spatial and temporal dependency of emission factors, especially in the tropics, may further refine future scenarios.

If climate policy is included, the upper end of the literature range decreases, since N_2O is directly targeted by climate policy. This is illustrated by the set of Representative Concentration Pathways that have the lowest N_2O emission for the most stringent climate policy scenarios. Emissions are reduced by about 30% in the RCP2.6 scenario compared to 2000 as a result of emission reduction measures, such as improved manure management systems [47,68].

Further emission reduction is limited by the fact that only a limited mitigation potential has been identified for N fertilizer use and animal husbandry [68]. This is partly related to the N cascade effect [69], where a reduction in one part of the cascade (e.g. ammonia emissions from stored animal manure) may cause an increased emissions in another part (N_2O emissions from manure spreading) [70]. An important aspect here is the recovery efficiency of N in food production, which determines the inputs of N fertilizer and animal manure in agricultural systems and the environmental losses (Figure 5). Especially in livestock production, recovery efficiencies are inherently low and can only be significantly increased by a shift from ruminant to white meat production [50]. Such projections are based on possible future technology development and improving management practices, which is a difficult task, particularly for low-income countries. Recovery efficiencies of 60–70% as currently achieved in some industrialized countries seem to be the maximum that can be achieved in practice, based on current knowledge and technology [71]. It should be noted that these recoveries refer to the production process only. The N that is finally consumed as food is much less [69].

In some cases, climate policy may actually increase N_2O emissions. For instance, both catalytic converters in cars and manure injections may increase N_2O emissions. A potentially more important factor is that N_2O emissions may increase as a result of the use of fertilizer in the production of bio-energy [72,73]. For first generation biofuel crops, the increase in N_2O emissions alone may offset the gains of using bio-energy.

Figure 5



Global recovery of N in crop and livestock production for 1970, 2000 and 2050 in industrialized and developing countries (IMAGE scenario for the International Assessment of Agricultural Science and Technology Development, [3]). Recovery is calculated as the N in the harvested parts divided by the input of fertilizer and manure (crops) or feed (livestock).

Projections for NH_3

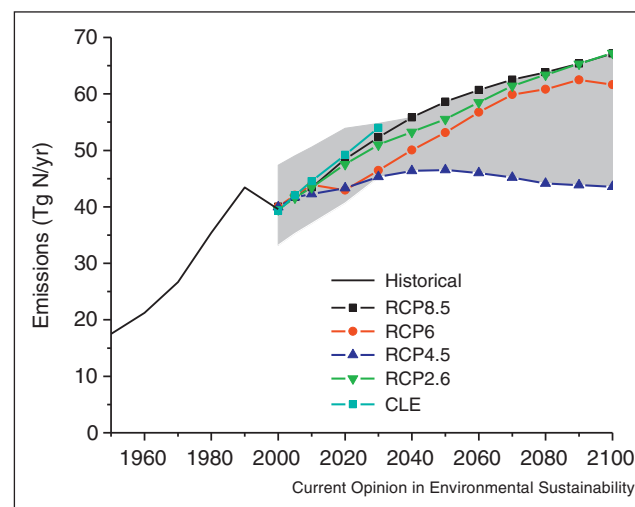
NH_3 emissions in 2000 amount to nearly 40 Tg N yr⁻¹. Most of these emissions originate from agricultural activities, that is, around 30 Tg yr⁻¹, mostly from animal husbandry [55]. About 10 Tg yr⁻¹ come from biomass burning (land use change and savanna burning). A minor part stems from traffic, an unwanted side-effect of three-way catalytic converters in, particularly, light-duty gasoline vehicles, contributing about 5% of total NH_3 emissions in the U.S.A. [74].

Expected growth of livestock production will lead to an increase of NH_3 emissions, while biomass burning is not expected to increase significantly in the future; in fact, emissions from biomass burning may even decrease as a result from slowing deforestation [75] (either autonomously or as a result of, for instance, measures in the context of climate policy). In other words, trends in future NH_3 emission mostly depend on agricultural practices and any measures that are introduced to decrease NH_3 emissions.

Not many scenarios deal with future NH_3 emissions. Dentener *et al.* [7] present a global scenario on the basis of the implementation of the SRES scenarios by the IMAGE model, showing an increase in the order of 40–50%. In the literature used for this paper, only the RCPs provide information of global trends in NH_3 emissions. Here, both the underlying baseline scenarios and the

RCPs themselves were assessed to compile the literature range. The set shows (as expected) that very little relationship exists between climate policy and NH_3 emissions, because NH_3 is not directly targeted by climate policy (Figure 6).

Figure 6



Future NH_3 emissions according to various scenarios (light grey area covers the 10–90th percentile; dark grey area the 25–75th percentile). Source: CLE [26] and RCP scenarios and the underlying baselines [40].

We note that that the set of four scenarios does not cover the full uncertainty range. Long-term projections of NH_3 emissions have only recently been constructed. For NH_3 from agriculture, the major global source, both the N recovery (see above) and the nutrient management in agricultural systems are important. Thus, the use and management of both N fertilizers and animal manure play a key role in NH_3 emissions [76]. Furthermore, there is a difference between systems dominated by grazing versus those where animals are confined. In the latter systems N can be lost from manure management systems, and also after spreading of stored manure in the field (the above-mentioned N cascade [69]). In scenarios this aspect is not always considered, but is crucial because NH_3 emission from intensive systems may exceed those from pastoral systems [50]. These emissions will also depend on the structure of future food demand, such as the consumption of meat from ruminants as compared to poultry.

Conclusions

Most scenarios containing information on future Nr gas emissions to the atmosphere imply that these emissions will slowly increase in the coming decades, followed by stabilization or decline. The Nr gas emissions have increased globally by about 150% in the 1950–2000 period. The assessment of scenarios published in the literature suggests that emissions may further increase in the 2000–2050 period, but at a much slower rate than in past. The 50% percentile of scenarios with and without climate policy ranges from about –5 to +35% increase compared to 2000. After 2050, emissions in most scenarios stabilize (scenarios without climate policy) or decline (scenarios with climate policy) leading to an overall range for the 50% percentile of –25% to +30% in 2100 compared 2000 (scenarios with and without climate policy).

Crucial factors that determine total future Nr emissions include the stringency of air pollution control measures in developing countries, agricultural development and future climate policy. The economic activities that drive Nr emissions are expected to continue to grow. This is especially the case for NO_x , which mostly originates from electric power production and transport (scenarios show growth rates for these activities in the order of 60 up to 300% in 2000–2050 period). N_2O emissions and NH_3 emissions are dominated by animal husbandry and N fertilizer use, activities which are expected to have lower growth rates (in the order of 50%). Nr emissions are not expected to grow at the same rate as the economic drivers: emission factors are expected to decline as a result of policy and structural changes. This is especially important for NO_x where at the global scale, declining emission factors alone may reduce emissions by up to 60% compared to the situation where emission factors would remain constant (but these depend on scenario assumptions). As N_2O and NH_3 emissions strongly depend on

agricultural activities, the development of the agricultural sector also is a key factor determining future emission growth (e.g. agricultural policies or the dietary shifts towards animal products). A third factor of crucial importance to both NO_x and N_2O emission is the introduction of climate policy. Such policies will directly (climate policies are likely to target N_2O emissions) or indirectly (NO_x emissions are influenced by the induced changes in the energy system) lead to emission reductions in most cases. However, some strategies may lead to increasing N_2O emissions, such as bio-energy production (to reduce CO_2 emissions) and strategies in animal manure management systems owing to the N cascade effect.

For making projections of future Nr gas emissions from agriculture, the most important tasks are to assess future recovery efficiencies of Nr in crop and livestock production systems, and to consider N cascade effects. Recovery efficiencies in livestock production are inherently low; changes in the share of ruminants versus that of pork and poultry appear to be more important for future Nr gas emissions than any management strategy. Recovery efficiencies in crop production are more readily increased, although in practice recoveries of 60–70% seem to be the maximum achievable. For NH_3 and N_2O projections, the N cascade involved in future development of intensive or industrial production systems of ruminants, and manure management systems is an important feature, which is often ignored in scenarios.

RCPs do not fully cover the range of possible futures. The representative emission scenarios (RCPs) were developed to explore climate change futures with and without stringent climate policy. Given their purpose, the RCPs are representative of the range of different emission trajectories for greenhouse gases (as shown here for N_2O). The scenarios also include a detailed set of emissions for air pollutants, given the role of these gases in climate change forcing. The overview here shows that the RCPs are not representative, however, of the full range of emissions scenarios of NO_x in the literature: the RCPs cover the low to medium range of the literature, while high emission scenarios are not well represented (on both sides the true uncertainty range might be even larger than suggested in our literature review). For the purpose of assessment of the impact of different air pollution control strategies it might therefore be useful to consider a wider range of scenarios. For NH_3 , present-day emission inventories are more uncertain than for NO_x . In addition, an assessment of a full range of future NH_3 emissions could not be made as a result of a lack of independently published scenarios.

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- of special interest
- of outstanding interest

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