



# The influence of water table depth on evapotranspiration in the Amazon arc of deforestation

John O'Connor<sup>1</sup>, Maria J. Santos<sup>2</sup>, Karin T. Rebel<sup>1</sup>, and Stefan C. Dekker<sup>1,3</sup>

<sup>1</sup>Copernicus Institute of Sustainable Development, Department Environmental Sciences, Utrecht University, Utrecht, the Netherlands

<sup>2</sup>University Research Priority Program in Global Change and Biodiversity and Department of Geography, University of Zürich, Zurich, Switzerland

<sup>3</sup>Faculty of Management, Science and Technology, Open University, Heerlen, the Netherlands

**Correspondence:** John O'Connor (j.c.oconnor@uu.nl)

Received: 30 January 2019 – Discussion started: 18 February 2019

Revised: 9 August 2019 – Accepted: 17 August 2019 – Published: 25 September 2019

**Abstract.** The Amazon rainforest evapotranspiration (ET) flux provides climate-regulating and moisture-provisioning ecosystem services through a moisture recycling system. The dense complex canopy and deep root system creates an optimum structure to provide large ET fluxes to the atmosphere, forming the source of precipitation. Extensive land use and land cover change (LULCC) from forest to agriculture in the arc of deforestation breaks this moisture recycling system. Crops such as soybean are planted in large homogeneous monocultures and the maximum rooting depth of these crops is far shallower than forest. This difference in rooting depth is key as forests can access deep soil moisture and show no signs of water stress during the dry season, while in contrast crops are highly seasonal with a growing season dependent on rainfall. As access to soil moisture is a limiting factor in vegetation growth, we hypothesised that if crops could access soil moisture, they would undergo less water stress and therefore would have higher evapotranspiration rates than crops which could not access soil moisture.

We combined remote-sensing data with modelled ground-water table depth (WTD) to assess whether vegetation in areas with a shallow WTD had higher ET than vegetation in deep WTD areas. We randomly selected areas of forest, savanna, and crop with deep and shallow WTD and examined whether they differ on MODIS Evapotranspiration (ET), Land Surface Temperature (LST), and Enhanced Vegetation Index (EVI), from 2001 to 2012, annually and during transition periods between the wet and dry seasons. As expected, we found no differences in ET, LST, and EVI for forest

vegetation between deep and shallow WTD, which because of their deep roots could access water and maintain evapotranspiration for moisture recycling during the entire year. We found significantly higher ET and lower LST in shallow WTD crop areas than in deep WTD during the dry season transition, suggesting that crops in deep WTD undergo higher water stress than crops in shallow WTD areas.

The differences found between crop in deep and shallow WTD, however, are of low significance with regards to the moisture recycling system, as the difference resulting from conversion of forest to crop has an overwhelming influence (ET in forest is  $\approx 2 \text{ mm d}^{-1}$  higher than that in crops) and has the strongest impact on energy balance and ET. However, access to water during the transition between wet and dry seasons may positively influence growing season length in crop areas.

## 1 Introduction

The Amazon rainforest has been reduced to 80% of its original size due to deforestation over the past few decades (Davidson et al., 2012). Land use and land cover change (LULCC) from forest and savanna to agricultural land disrupts the Amazonian water cycle due to changes in evapotranspiration, infiltration, and runoff (Fearnside, 1997; Lawrence and Vandecar, 2014). Changes in evapotranspiration result in major changes to the water energy balance, as forest vegetation has high evapotranspiration rates and is re-

placed with agricultural vegetation with lower evapotranspiration, which results in a lower latent heat flux and higher sensible heat flux (Swann et al., 2015). In addition, a decline in evapotranspiration reduces the available atmospheric moisture, which can reduce rainfall.

Differences in vegetation structure are suggested to be the main drivers affecting the evapotranspiration rates. Three major land cover classes can be identified at the Amazon arc of deforestation; forest, savanna (Brazilian cerrado; here we use “savanna” to keep terms equivalent to the land cover classification used), and agriculture. Forest vegetation has the highest total leaf surface area, while savanna has a lower leaf surface area owing to its mixed structure of grasses, shrubs, and trees with a more open canopy, and agricultural vegetation usually has a lower leaf area (Asner et al., 2003; Costa et al., 2007). This difference in leaf area lowers the potential surface area for both interception evaporation and transpiration. In addition, the rooting depth of forest savanna and agricultural vegetation differs greatly (Costa and Foley, 2000). Forest vegetation has deep roots which facilitate access to deep soil moisture, maintaining their supply of water necessary for photosynthesis even during the dry season. Therefore, forest evapotranspiration remains high throughout the year, unaffected by periods of low rainfall (Maeda et al., 2017; Staal et al., 2018). While the rooting depths of savanna tree species have been shown to be deep, the savanna landscape also contains more open shallow rooted shrubs and grasses. Following LULCC from forest or savanna to agriculture, the new vegetation cover lacks deep roots and therefore no longer accesses deeper soil moisture. Over the past few decades, the developing agricultural industry driven by international demand encouraged extensive LULCC (Brando et al., 2014; Foley et al., 2007; Sampaio et al., 2007) concentrated along the southern and eastern edges of the Amazon in an area known as the arc of deforestation (Costa and Pires, 2010; Malhi et al., 2008). LULCC negatively impacts the ecosystem service provision of the Amazon, including highly valuable services such as carbon storage and sequestration and moisture recycling and regulation. However, little is known about whether LULCC that occurred in areas with a shallow water table depth (WTD) facilitates access to water and leads to higher vegetation productivity and evapotranspiration compared to areas with a deep WTD. Understanding the effect that LULCC has on evapotranspiration is important as the loss of evapotranspiration impacts both climate and precipitation on local and regional scales.

Local climate can be impacted by LULCC due to changes in the energy balance as loss of evapotranspiration reduces latent heat and increases sensible heat. Studies in the Amazon have shown that temperatures increase on average by 1.4 °C, with a maximum of 7 °C following conversion to crop (Badger and Dirmeyer, 2015). The seasonal impact of LULCC is particularly strong during the dry season as crop evapotranspiration is at its lowest, latent heat flux can be reduced by 78 %, and the sensible heat flux can increase by 85 % rel-

ative to forest (Ponte De Souza et al., 2011). The loss of evapotranspiration impacts rainfall both locally and on the continental scale. Evapotranspiration returns water to the atmosphere, where it can precipitate again either in situ or be carried further downwind (Eltahir and Bras, 1994a). Large forests like the Amazon, because of their density and extent, create large evapotranspiration fluxes, leading to underpressure over land, and the pressure differences draw moisture towards land (Makarieva and Gorshkov, 2007; Sheil, 2014). As much as 70 % of rainfall in the Amazon and southern Brazil is a result of Amazonian evapotranspiration (van der Ent and Savenije, 2011). This evapotranspiration precipitation cycle is highly important in both maintaining the forest itself and providing precipitation to non-forested areas. LULCC reduces the evapotranspiration and breaks this moisture recycling system, resulting in lower rainfall locally and downwind. The seasonal loss of evapotranspiration in crop areas during the dry season is of great significance: evidence already suggests that LULCC has resulted in a lengthening of the dry season (Costa and Pires, 2010; Debortoli et al., 2017). Model simulations predict that if deforestation continues, by 2050 the loss of evapotranspiration will result in a negative effect further reducing forest cover and evapotranspiration (Foley et al., 2007; Spracklen et al., 2012). The conversion of forest and savanna to agricultural land in Brazil is driven by an increasing demand for agricultural production, which has almost doubled since 2000 (Zalles et al., 2019); however, losses in evapotranspiration could lead to subsequent losses in agricultural productivity as rainfall is reduced and the growing season is shortened (Oliveira et al., 2013).

Crops in the Amazon arc of deforestation are predominantly rainfed and as such impacted by the high seasonality in rainfall unseen in forest vegetation. Forest vegetation provides an optimum structure for evapotranspiration due to its tall complex dense canopy and deep root systems which can access deep soil moisture stores and maintain high transpiration rates even during periods of low rainfall (Nepstad et al., 1994; Sheil, 2014). Savanna has a mixed composition, with both trees and grass layers, and more open canopy and a lower leaf area. Savanna trees can have a deep rooting depth (> 10 m) facilitating access to deep soil water (Canadell et al., 1996). Agricultural crops are known to contribute much less to evapotranspiration as a result of their shorter canopy and simpler structure (Fearnside, 1997). In addition, agricultural crops lack the deep root systems of forest which are credited with maintaining evapotranspiration throughout the dry season (Nepstad et al., 1994). In theory, if vegetation continues to access the water table within the root zone, then this vegetation will continue to transpire during periods of reduced rainfall. Thus limited access to soil moisture is an important limiting factor for photosynthesis and transpiration. Shallow water table depths across South America are widely distributed and correspond to an area of approximately 36 % of the Amazon (Fan and Miguez-Macho, 2010). We hypothesise that areas of shallow WTD allow shallow

rooted vegetation to access soil moisture, potentially facilitating vegetation productivity and higher evapotranspiration when compared to areas of deep WTD. Experimental manipulation of WTD using sub-irrigation systems of soybean demonstrated that shallow WTD benefitted productivity and increased yield (Kahlow et al., 2005; Mejia et al., 2000). In the Amazonian arc of deforestation, irrigation of crops is relatively uncommon (Lathuillière et al., 2012), and increases in agricultural productivity have been achieved primarily by increasing the area of crops (Oliveira et al., 2013). If agricultural vegetation can access soil moisture in these shallow WTD areas, it could potentially increase the growing season length and productivity without the need for investment in irrigation systems. In turn, less land would be required to achieve the same agricultural output. During the wet season, soybean can reach rates of evapotranspiration similar to that of forest (Costa and Foley, 2000). Some studies have suggested that the difference in annual ET between forest and agricultural crops is primarily due to access to water during the dry season (Costa et al., 2007).

In this study, we use a number of freely available remote-sensing products in combination with modelled water table depth to investigate whether naturally occurring shallow water table depth could increase evapotranspiration compared to deep water table depth. We expect the greatest influence to be seen in crop areas as they have the shallowest rooting depth and are most dependent on precipitation. As reported in other studies the influence of WTD should not be visible for deep rooted vegetation (Nepstad et al., 1994) like forest and some savanna species. As savanna has mixed vegetation and rooting depths, we expected to find some differences in ET as a result of deep and shallow WTD. We expect that the differences as a result of WTD will be greater in the transition periods between wet and dry seasons as rainfall as a water source is limited. In areas of shallow WTD, the saturated zone is closer to the root zone of vegetation. In these locations we, therefore, expect crop vegetation to be buffered against the reduction in rainfall during the dry season transition and experience drought conditions later, thus delaying the decline of transpiration due to the dry season. Similarly, during the wet season transition, we expect that areas of shallow WTD will have higher productivity as crop vegetation may access the shallow WTD to supplement their demand when rainfall is low, therefore growing sooner than areas with deep WTD, effectively shortening the dry season. Finally, we discuss whether differences found in ET between deep and shallow WTD are important for moisture recycling and vegetation productivity, and what the implications for future LULCC are.

## 2 Methods

### 2.1 Study area

The study area is located in the southern Amazon, mostly in the northern region of Mato Grosso and incorporating the border area with Pará (Fig. 1). Mato Grosso is classified into three major biomes, with rainforest in the north, cerrado (a vegetation type that resembles savanna) in the central region, and wetlands in the southwest (Kastens et al., 2017; Lathuillière et al., 2012). The climate has two seasons, the wet season in the austral winter and the dry season in austral summer; the dry season lasts around 5 months with an annual average rainfall of 2000 mm and monthly mean temperatures between 22 and 26 °C (Arvor et al., 2014). This precipitation level is within the natural range supporting both savanna (700 to 2000 mm yr<sup>-1</sup>) and forest (1000–2500 mm yr<sup>-1</sup>). Mean elevation over the study area is 345 m ± 100 m, with a maximum of 700 m and a minimum of 100 m. Runoff in the Amazon basin is usually low, with groundwater convergence accounting for as much as 90 % of streamflow (Miguez-Macho and Fan, 2012). The mean WTD of the study area is 12 m, with approximately 20 % shallow (< 2 m). The maximum WTD is 60–70 m. This region is well known as a dynamic agricultural frontier – the arc of deforestation – with high rates of LULCC, where forest and savanna are converted for extensive agriculture, mostly cattle ranching and soy production (Kastens et al., 2017). Mato Grosso is the leading producer of agricultural crops such as soybean in Brazil (Gusso et al., 2014). We chose a 750 km × 750 km study area which is centrally located in the arc of deforestation and has large areas of primary forest (73 %), savanna (19 %), and crops (3 %).

### 2.2 Datasets

#### 2.2.1 Remote-sensing data

Remote sensing offers excellent tools for monitoring changes in vegetation over large regions as it provides full geographic coverage and high temporal frequency at spatial scales relevant to most Earth system processes (Chambers et al., 2007). Here we use three separate products from the Moderate Resolution Imaging Spectrometer (MODIS), namely MODIS Evapotranspiration (MOD16A2), MODIS Land Surface Temperature (MOD11A2), and MODIS Enhanced Vegetation Index (MOD13A2), to assess the influence of WTD on evapotranspiration. MODIS remote-sensing products were used as they offer a moderate spatial resolution and a high temporal resolution, which is ideal for examination of seasonal processes. We chose to perform the analysis for the currently available MODIS land cover archive using data from 2001 to 2012. In addition, this period represents a time with high variability of precipitation extremes in which the Amazon experienced droughts and floods, and it could



**Figure 1.** Study area of the arc of deforestation of the Amazon, in northern Mato Grosso. The inlaid image shows the MODIS land cover classification map (2001) for the three land cover classes analysed. Forest – green, Savanna – beige, Crop – yellow, and Other – grey. Due to the sinusoidal projection of MODIS satellite data, the study area looks distorted. Satellite image data: © Google Earth, Landsat/Copernicus.

depict the variability the system experiences (Nobre et al., 2016a). Data were downloaded from the NASA data sharing portal (<https://earthdata.nasa.gov/>, last access: April 2018). Data were rescaled to 1 km resolution; no additional post-processing was conducted.

MODIS Evapotranspiration (hereafter ET) data (Mu et al., 2011) provide 8 d accumulated evapotranspiration at 500 m resolution (rescaled to 1 km). The ET dataset is one of the best available datasets due to its relatively high spatial and temporal resolution; as such, it has been widely used to investigate the effect of land use change on evapotranspiration in the Amazon (Loarie et al., 2011; Neill et al., 2013; Vergopolan and Fisher, 2016). The baseline algorithm for the MODIS ET product is based on the Penman–Monteith equation and combines parameters such as land cover, leaf area index (LAI), albedo, and fraction of photosynthetically active radiation (FPAR) directly observed with or modelled from MODIS data, with reanalysis data on radiation, air temperature, and humidity from the Global Modelling and Assimilation Office (Mu et al., 2011). The MODIS ET products were previously tested over the Amazon by comparing their outputs with eddy-covariance tower data, showing that the product is more accurate over longer temporal scales (monthly time steps) and larger areas (e.g. drainage basin) (Ruhoff et al., 2013; Velpuri et al., 2013). While the MODIS ET product is known to be underperforming at fine temporal resolutions and newer novel methods show promising results at nine flux sites across the Amazon (Xu et al., 2019), we believe that the application of the new method for our question on the influence of WTD and our time series analysis was

beyond the scope of this study. This is also the reason why we chose to also analyse the effects of WTD on satellite-retrieved EVI and LST. As with these additional products, differences might be detectable and potentially indicate the effects of WTD on the water cycle.

MODIS Land Surface Temperature (hereafter LST) provides an 8 d mean daytime land surface temperature in Kelvin at 1 km resolution. LST data are produced by detection of thermal infrared radiation between 3 and 15  $\mu\text{m}$  spread across 15 bands of the thermal sensor onboard the MODIS satellite system and temperatures are modelled based on land cover classification with a clear-sky accuracy of 1  $^{\circ}\text{K}$  (Wan, 2014). MODIS LST data were converted to degrees Celsius. Evapotranspiration in the Amazon has been shown to result in a net cooling effect (Bonan, 2008); therefore, areas with lower LST will be observed in areas of higher ET (Eltahir and Bras, 1994b).

MODIS Enhanced Vegetation Index (hereafter EVI) provides an observation of vegetation greenness at a frequency of 16 d and 500 m resolution (rescaled to 1 km). EVI is a vegetation index that measures greenness as a proxy for productivity (Huete et al., 2002). It was developed to improve upon the normalised difference vegetation index (NDVI), as EVI is less sensitive to saturation in highly dense canopies such as those in the Amazon, and EVI also corrects for canopy background effects and atmospheric aerosol effects (Huete et al., 2002). This MODIS product offers an observation of vegetation productivity as it measures “greenness” and is correlated with photosynthesis/evapotranspiration (Mu et al., 2011). Thus vegetation with adequate access to water near

the root zone will have a comparatively higher EVI than vegetation which is water stressed. This higher EVI, in turn, would correspond to areas of higher ET.

In addition, we also used the MODIS land cover product for selection of our analysis sites (see below). MODIS land cover (hereafter land cover) provides a classification of global land cover at 1 km resolution, and it is annually updated. For this study, we only used pixels that were classified as the same land cover type during the entire study period 2001–2012. The study area chosen provides a sufficient number of representative pixels for random selection of each land cover type. The use of stable land cover classes was necessary to determine and describe the patterns of ET, LST, and EVI over time and assess the effects of WTD on such trends without the confounding effect of land cover change. Further, we used MODIS land cover as it is the same land cover classification map as used for the MODIS ET product (Friedl et al., 2010) to avoid effects of land cover classification errors from different maps.

Over the Amazon cloud cover and shadows are an issue, especially in the wet season. Pixels with high cloud cover were excluded from the analysis. The high seasonal difference in cloud cover is clear; at each time step we used a spatial mean of only available pixels, and due to our large sample size we still have enough pixels for the analysis (see Fig. S10.1 in the Supplement). We compared the cloud cover per land cover class and found no bias or significant differences between deep and shallow areas.

Topography might influence the MODIS data in a number of ways. Elevation can influence meteorological forcing (i.e. temperature and vapour pressure), which is used to calculate ET. Topography can also influence water availability on a pixel due to the slope and catchment size of the surrounding area, impacting water available to vegetation, therefore influencing ET and EVI. Serious errors due to topography are filtered by the MODIS quality control dataset, and these pixels were excluded from our analysis. We used SRTM (Shuttle Radar Topography Mission) data to examine elevation and calculate the topographic wetness index (an integrated measure of water accumulation) of our studied pixels. No significant differences were found between the elevation of deep and shallow WTD areas of forest or savanna. Crop elevation was found to be significantly different between deep and shallow areas for half of our randomisations. However, the difference in mean elevation was only 10 m, leading us to believe that this will not have a strong impact on the meteorological forcing data or ET. We found no significant differences in the topographic wetness index between deep and shallow land covers (see Fig. S9.4).

Finally, the Tropical Rainfall Measuring Mission (hereafter TRMM) 3B42 provides daily precipitation at a resolution of 0.25° (downloaded from <https://earthdata.nasa.gov/>, last access: April 2018). We calculated daily mean rainfall of our study area using the TRMM data, which were then used to calculate the seasonality of rainfall, i.e. start of the

dry season and the wet season across the study area, and not per pixel (see below for further details).

### 2.2.2 Water table depth

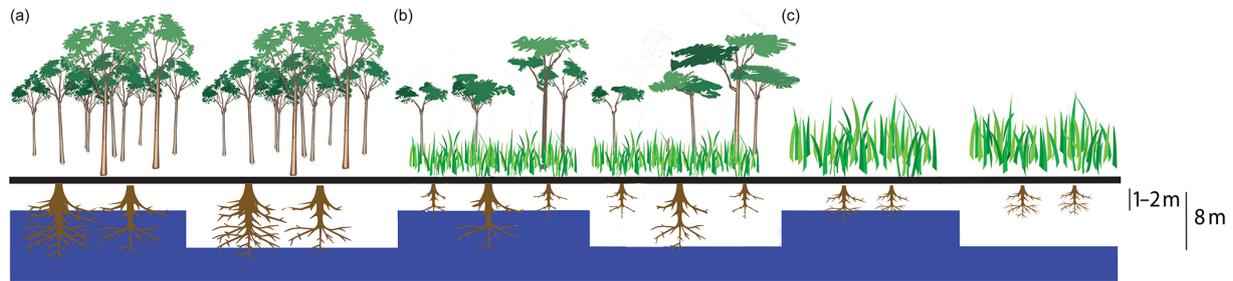
WTD values were extracted from the Fan and Miguez-Macho (2010) equilibrium WTD model of South America at 30 arcsec ( $\sim 1$  km). The model was created as a long-term mean water table depth using a combination of literature-reported depths and national databases of groundwater table depth, most of which are from drinking water wells from areas of high population. These data are interpolated using a groundwater model forced by present-day climate, terrain, and sea level. We used the output of the model to obtain WTD data, which were projected to the same sinusoidal projection of the MODIS data. The equilibrium WTD model is intended for use in dynamic simulations, and although our study is not the intended use of the WTD model, it is the best currently available. As the WTD model output is in “equilibrium”, it gives a better indication of the annual average WTD compared to interpolated WTD measurements, which may be biased depending on when they were recorded. The authors compared their WTD calculations with values reported in the literature and found good agreement for shallower WTD; however, the model overestimated deep WTD. We selected two broad WTD classes in order to further reduce some of the uncertainty around this key parameter: shallow  $< 2$  m and deep  $> 8$  m (and we will refer to these as such hereafter). Figure 2 shows a theoretical graphical representation of the difference between forest (deep rooting depth), savanna (mixed rooting depth), and crop (shallow rooting depth) land cover classes. These depths were selected as they represent rooting depth values for crop and forest vegetation from the literature (Fan et al., 2016; Moreira et al., 2000; Nepstad et al., 1994; Setiyono et al., 2008).

## 2.3 Sampling design

### 2.3.1 Spatial sampling

We chose to avoid pixels which experienced LULCC during the study period as we wanted to use the full time series for each pixel. We used MODIS land cover to identify pixels of each land cover class which remained unchanged between years and used these for analysis. We combined three land cover classes with the two water table depths and analysed the following classes: Forest Deep, Forest Shallow, Savanna Deep, Savanna Shallow, Crop Deep, and Crop Shallow.

For each class, we randomly selected 1000 pixels and performed this random selection 20 times to account for the effect of the randomisation process on the results. This random selection method increased computational efficiency by limiting the number of total pixels examined and producing comparable group sizes for statistical analysis. During the wet season the number of usable pixels was as low as



**Figure 2.** Diagram showing that forest (a) root depth can reach until the saturated zone in both shallow (< 2 m) and deep (> 8 m) WTD, savanna (b) has a mixed rooting depth, with only tree roots reaching deep WTD, and crop (c) vegetation has a low maximum rooting depth – crops having a maximum rooting depth of 2 m and savanna having a maximum rooting depth > 10 m (Canadell et al., 1996). Shallow roots can reach the saturated zone in shallow WTD (< 2 m); however, they cannot reach the saturated zone in deep WTD (> 8 m).

200–300 pixels per class for some time steps, while in the dry season the number of usable pixels was above 900 (see Fig. S8.1).

### 2.3.2 Data analysis

The Amazon arc of deforestation is located in a region that has two major seasons defined by the difference in rainfall, the wet season from October to March (approximately 1500 mm) and the dry season from April to September (approximately 400 mm). The difference in rainfall can have significant impacts as the area can be prone to both seasonal flooding and droughts. In recent years the Amazon arc of deforestation has undergone an increased frequency of extreme weather events, with drought in 2005 and 2010 and flooding in 2009 and 2012 (Nobre et al., 2016b). These extreme climatic conditions can have a large influence on ET, and vegetation distribution as waterlogging of soils can lead to anoxia in the root zone. Due to the selection of only consistently classified pixels the influence of waterlogging can be avoided as over time these areas will fall under different classifications. Investigation into the drivers of these extreme variations and how each land cover class is influenced is however beyond the scope of this study.

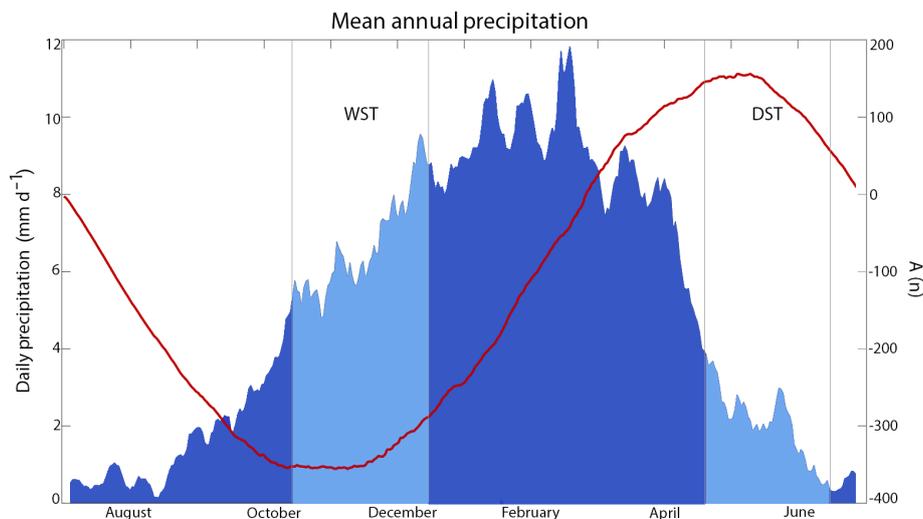
Analysis of the data was conducted using three primary time periods. We compared mean daily values of ET, EVI, and LST between deep and shallow WTD as this gives an indication of the influence of WTD on our land cover classes without considering the seasonal variation. We then compared ET, EVI, and LST of our land cover classes during the dry season transition (DST) and wet season transition (WST) periods.

For each year we calculated the DST and WST using mean daily precipitation of our study area from TRMM with the anomalous accumulation method of Liebmann et al. (2007). This method uses the following equation:

$$A(n) = \sum_{n=1}^{\text{day}} [R(n) - \bar{R}], \quad (1)$$

where  $R(n)$  is daily precipitation and  $\bar{R}$  is the average daily precipitation. Calculation of the anomalous accumulation begins in the driest month of the year, when the difference between daily precipitation and annual average is summed to a running total of the anomalous accumulation ( $A$ ). The wet season onset is defined as the beginning of the longest period where the anomalous accumulation remains positive, while the dry season onset is defined as the day after this anomalous accumulation reaches its maximum (Fig. 3). These onset points of the dry and wet seasons were applied to find the closest time stamp from each MODIS product in the time series. We then considered the DST to last on average eight repeats in the MODIS record (five for EVI due to the lower frequency of the product) and the WST seven repeats (four for EVI). We used an average value for each remote-sensing product over these transition periods to assess the difference between shallow and deep WTD on evapotranspiration.

The WST and DST periods selected as LULCC in the arc of deforestation are correlated with a lengthening of the dry season, in particular delays in the WST (Butt et al., 2011; Dubreuil et al., 2012; Fu et al., 2013). Recently, evapotranspiration has been shown to draw moist air over the Amazon, triggering the wet season before migration of the ICTZ (Wright et al., 2017). In this study, we focus our analysis on differences in the DST and the WST. During the DST, there is already significant drydown (anomalous accumulation is at a maximum, and precipitation already went down before; see Fig. 3), which should be apparent in vegetation without access to deeper water sources. Further into the dry season, other factors may cause a decline in transpiration as well, like heat stress. During the WST, we focus on the recovery of the vegetation, which should be faster when they have access to deeper water sources, like deep roots or a shallow WTD. Thus shallow rooted vegetation in shallow WTD areas may have higher access to water as their root zone is closer to the water table. This will likely produce higher ET and EVI and lower LST during the DST than shallow rooted vegetation in deep WTD areas. This is because the WTD is much deeper and further from the vegetation rooting zone, which leads to



**Figure 3.** Mean annual precipitation of the study area calculated from TRMM with a 7 d average for graphical smoothing. Wet season transition (WST) and dry season transition (DST) periods are represented in lighter blue. The vertical lines represent the average start and end dates; however, exact dates were calculated per year between July 2001 and July 2012. The red line represents the anomalous accumulation method  $A(n)$  from Liebmann et al. (2007).

a lack of access to water, and the vegetation will likely be stressed. Similarly, during the WST, shallow rooted vegetation in shallow WTD may exhibit higher ET and EVI and lower LST than that in deep WTD because vegetation cannot yet be sustained by precipitation alone. We do not expect these differences with deeply rooted vegetation.

We tested whether ET, LST, and EVI followed a normal distribution using the Kolmogorov–Smirnov test. This test served two purposes, to assess whether parametric statistics could be used and also to indicate whether the WTD influences the frequency distribution of ET, LST, and EVI. Since a large number of response variables were not normally distributed, we chose to use non-parametric methods. Therefore, a Wilcoxon rank sum test was used to test whether there was a significant difference in median ET, LST, and EVI due to the deep and shallow water table.

We further examined the frequency distribution of deep and shallow WTD of each of the datasets using the methodology of Wilcox (2012), where the lower and upper quantiles of the distribution are compared. Wilcox's method utilises bootstrapping in order to compare the distribution of the 10th and 90th quantiles using the Wilcoxon–Mann–Whitney test. Due to our large sample size, 100 bootstrapped datasets were used.

Statistical analysis between each deep and shallow land cover pair was performed separately each year for all 20 randomisations; e.g. differences in forest ET were tested for significance in  $12 \text{ yr} \times 20$  randomisations. For 1 year, the difference in ET, EVI, or LST was considered statistically significant when more than 66.7% of the randomisations were significant and an overall significance was determined if the majority (> 50%) of the years were significant. Statis-

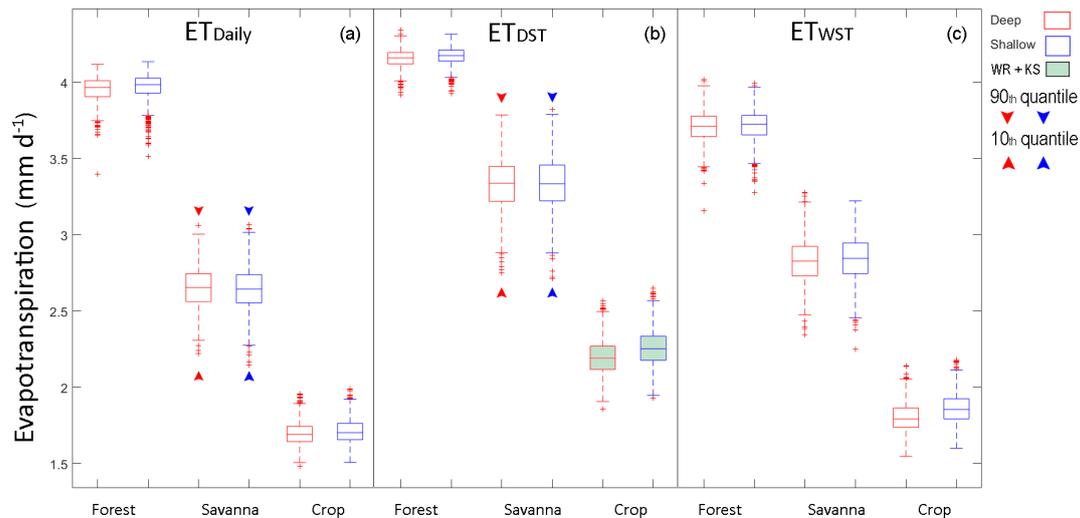
tical analysis was performed using the Matlab R2018a (The MathWorks Inc., Natick, USA) statistical toolbox and the Wilcox (2012) quantile distribution tool.

### 3 Results

The following results section is split into three subsections, one for each of the MODIS products used in the analysis. Each of the subsections and accompanying figures follows the same structure. Each figure uses three panels for the three time periods on the analysis: (a) annual daily mean, (b) daily mean during DST, and (c) daily mean during WST. Each panel has three pairs of box plots which represent the deep and shallow WTD data for forest, savanna, and crop.

#### 3.1 Effect of groundwater depth on evapotranspiration

None of the three land cover classes had significant differences in the average daily evapotranspiration ( $ET_{\text{daily}}$ ) between deep and shallow WTD areas (Fig. 4a). However, while we did not find consistent significant differences, in both forest and crop  $ET_{\text{daily}}$  we do see a trend towards higher  $ET_{\text{daily}}$  in shallow WTD areas for both (average  $\pm$  standard deviation: Forest Deep =  $3.953 \pm 0.08 \text{ mm d}^{-1}$ , Forest Shallow =  $3.967 \pm 0.09 \text{ mm d}^{-1}$ ; Crop Deep =  $1.697 \pm 0.07 \text{ mm d}^{-1}$ , Crop Shallow =  $1.713 \pm 0.08 \text{ mm d}^{-1}$ ). Interestingly, we found significant differences for savanna at the extremes of the distributions, depicted by the arrows in Fig. 4. Both the 10th and 90th quantiles of  $ET_{\text{daily}}$  were significantly higher in deep WTD areas than in shallow ones (difference of



**Figure 4.** (a) Average daily evapotranspiration (ET) annually  $ET_{\text{daily}}$ , (b) during the dry season transition period  $ET_{\text{DST}}$ , and (c) during the wet season transition  $ET_{\text{WST}}$ . Red boxes represent deep WTD. Blue boxes represent shallow WTD. Significant results are shown by the green filled boxes if significance was found with both Wilcoxon rank (WR) and Kolmogorov–Smirnov (KS). Significant differences in the 10th and 90th quantiles are depicted by the arrows.

10th =  $0.017 \text{ mm d}^{-1}$ , difference of 90th =  $0.02 \text{ mm d}^{-1}$ ; see Table S2.4 in the Supplement for all the quantile analyses).

Clear differences in seasonality occur between the different land cover types (see Figs. S3.1–S3.3). During the wet season mean ET of all land cover types can be above  $4 \text{ mm d}^{-1}$ . Both crop and savanna show clear suppression of ET during the dry season.

Crop ET during the DST (hereafter  $ET_{\text{DST}}$ ) was significantly higher in shallow than deep WTD areas (average  $\pm$  standard deviation ET: Deep =  $2.196 \pm 0.11 \text{ mm d}^{-1}$ , Shallow =  $2.26 \pm 0.12 \text{ mm d}^{-1}$ ; see the green filled boxes in Fig. 4b). Again we observed significant differences at the extremes of the distribution for savanna: on average the 10th quantile of  $ET_{\text{DST}}$  was higher in shallow (average difference =  $0.003 \text{ mm d}^{-1}$ ) and on average the 90th quantile of  $ET_{\text{DST}}$  was higher in shallow (average difference =  $0.005 \text{ mm d}^{-1}$ ).

ET during the WST (hereafter  $ET_{\text{WST}}$ ): while on average  $ET_{\text{WST}}$  was higher in shallow WTD areas than in deep WTD areas (average difference: Forest =  $0.01 \text{ mm d}^{-1}$ ; Savanna =  $0.01 \text{ mm d}^{-1}$ ; Crop =  $0.06 \text{ mm d}^{-1}$ ), this difference was not significant (Fig. 4c).

### 3.2 Effect of groundwater depth on land surface temperature

We found that the distribution of the average land surface temperature ( $LST_{\text{daily}}$ ) was significantly different only for savanna and the 90th quantile of crop. Deep WTD areas of savanna showed a distribution skewed towards lower temperatures (average  $\pm$  standard deviation LST: Deep =  $31.705 \text{ }^\circ\text{C} \pm 0.38$ , Shallow =  $31.848 \text{ }^\circ\text{C} \pm 0.37$ ); see

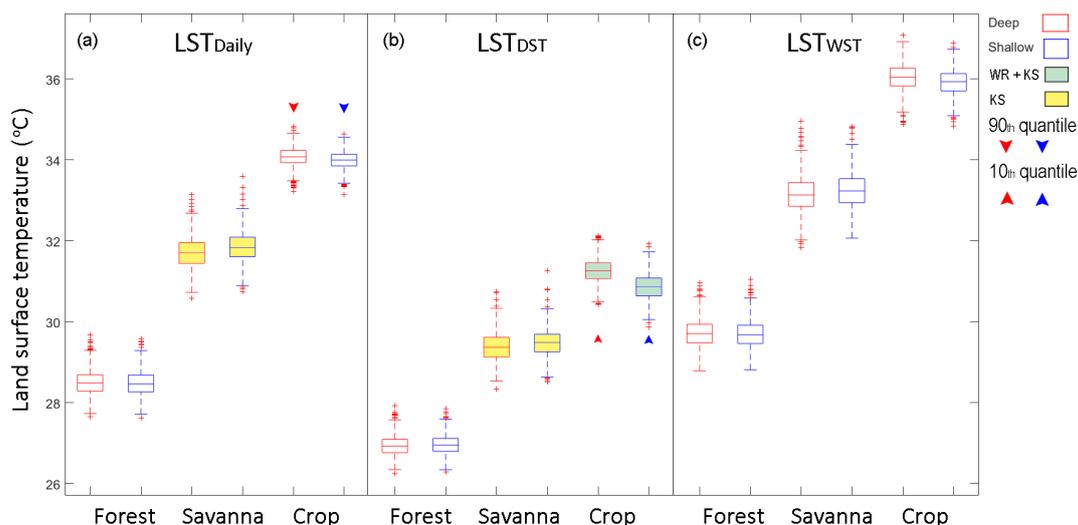
yellow filled boxes in Fig. 5a. The 90th quantile of crop  $LST_{\text{daily}}$  deep WTD areas was on average  $0.1 \text{ }^\circ\text{C}$  higher than in shallow WTD areas. Although this is only part of the distribution, it indicates that the warmest crop areas are found in deep WTD.

LST shows clear seasonal differences between the different land covers. Crop LST has the highest range in LST, with the warmest period coming towards the end of the dry season (August/September) (Figs. S5.1–S5.3). During the DST, we found that crop in deep WTD areas had a significantly higher LST than in shallow WTD areas (average  $\pm$  standard deviation LST: Deep =  $31.256 \pm 0.29 \text{ }^\circ\text{C}$ , Shallow =  $30.864 \pm 0.31 \text{ }^\circ\text{C}$ , green filled boxes in Fig. 5b). In addition, the 10th quantile of the crop distributions was significantly higher, by  $0.42 \text{ }^\circ\text{C}$  in deep WTD areas than in shallow ones. During these periods we found again a significant difference in the distribution of savanna, where deep savanna distribution was skewed towards lower LST values. No significant differences were found during the WST (Fig. 5c).

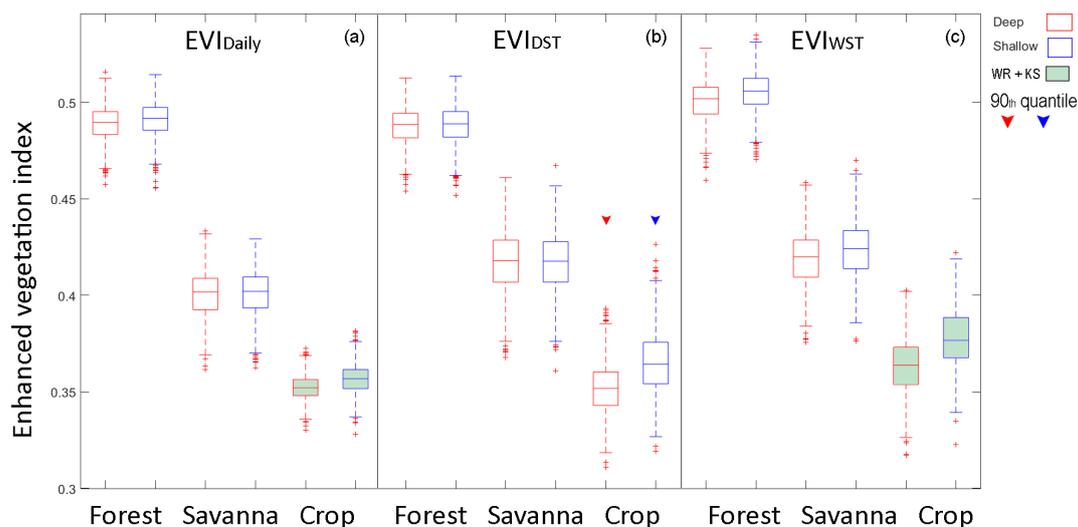
### 3.3 Effect of groundwater depth on enhanced vegetation index

We found significant differences in daily average EVI ( $EVI_{\text{daily}}$ ) between deep and shallow WTD only in crop (average  $\pm$  standard deviation EVI: Deep =  $0.352 \pm 0.01$ ; Shallow =  $0.357 \pm 0.01$ ), with shallow WTD areas' EVI being higher than that of deep WTD areas (Fig. 6a, green filled boxes).

Seasonality in EVI is shown in Figs. S7.1–S7.3. Crop EVI shows the highest variation among land cover types. When looking at the DST (May/June) of crop EVI it seems that



**Figure 5.** (a) Average daily land surface temperature (LST) annually  $LST_{daily}$ , (b) during the dry season transition period  $LST_{DST}$ , and (c) during the wet season transition  $LST_{WST}$ . Red boxes represent deep WTD. Blue boxes represent shallow WTD. Yellow filled boxes represent a statistical difference in skewness, calculated by Kolmogorov–Smirnov, and green filled boxes represent statistical differences by both Wilcoxon rank and Kolmogorov–Smirnov. Significant differences in the 10th and 90th quantiles are depicted by the arrows.



**Figure 6.** (a) Average enhanced vegetation index (EVI) annually  $EVI_{daily}$ , (b) during the dry season transition period  $EVI_{DST}$ , and (c) during the wet season transition  $EVI_{WST}$ . Red boxes represent deep WTD. Blue boxes represent shallow WTD. Green filled boxes represent statistical differences by both Wilcoxon-rank and Kolmogorov–Smirnov tests. Significant differences in the 90th quantile are depicted by the arrows.

the response is delayed in shallow WTD compared to deep WTD; for the WST (October/November) it seems that EVI in shallow areas increases faster than in deep WTD areas.

Mean EVI during the DST ( $EVI_{DST}$ ) for crop showed a trend towards higher EVI in shallow WTD areas; however, this difference was only significant in 5 of the 11 years, and therefore is not considered consistent enough to be statistically significant (average  $\pm$  standard deviation EVI: Deep =  $0.352 \pm 0.01$ , Shallow =  $0.3656 \pm 0.01$ ; Fig. 6b, Table S6.8). The 90th quantile EVI of crop was signifi-

cantly higher in shallow WTD areas than deep. During the WST ( $EVI_{WST}$ ), crop was the only different class where EVI was significantly higher in shallow WTD areas than in deep WTD areas (average  $\pm$  standard deviation EVI: Deep =  $0.364 \pm 0.01$ , Shallow =  $0.378 \pm 0.02$ , green filled boxes in Fig. 6c).

#### 4 Discussion

In this study, we tested the hypothesis that areas of shallow water table depth (WTD) would have higher evapotranspiration when compared to areas of deep WTD. As crop vegetation has the shallowest roots (< 2 m), we expect to see the largest influence of WTD in crop vegetation. In areas of deep WTD the root zone is far from the saturated zone, resulting in less uptake of deep soil water, while in areas of shallow WTD the root zone is close to the saturated zone, therefore providing the crops with access to groundwater. However, we found no support for this as the annual daily mean ET was not different between crop in deep and shallow WTD areas. One potential explanation is that since crops experience high seasonality, this annual variability may override differences between deep and shallow WTD areas in the daily average values of ET. For example, average crop ET reaches a maximum of  $3.5 \text{ mm d}^{-1}$  in the wet season, while the dry season ET reaches a minimum of  $0.4 \text{ mm d}^{-1}$ . Interestingly we found significant differences in annual mean LST and EVI for crop. For LST, we found that the upper 90th quantile was  $0.11 \text{ }^\circ\text{C}$  higher in deep than in shallow WTD areas. While this difference is only found in the 90th quantile of the distribution, it does indicate that LST in deep WTD areas can reach higher temperatures than shallow WTD areas. In addition, we found that crops in shallow WTD areas had a significantly higher EVI than in deep WTD. Crop EVI in shallow WTD areas is 1.2 % higher than in deep WTD. This provides support to our hypothesis that crop would have higher EVI in shallow WTD compared to deep WTD areas. The maximum rooting depth for most crops in the region is 2 m; in shallow WTD areas this means the root zone is close to the WTD and would have access to water, while in deep WTD the roots are far from the saturated zone. This access to water in shallow WTD areas could also lead to higher ET, and therefore evaporative cooling could explain the cooler temperatures in the 90th quantile.

The second part of our hypothesis was that the effect of WTD would be most evident during the transition periods between wet and dry seasons when rainfall is reduced and vegetation activity is limited by access to soil moisture. We found support for this hypothesis during the DST. In the DST, crop ET was significantly higher in shallow WTD areas and crop LST was significantly lower in shallow WTD areas, while in crop EVI we saw a trend towards higher EVI in shallow WTD areas (significant differences were only found in 5 of the 11 years). While the difference in crop ET is not large ( $0.063 \text{ mm d}^{-1}$ , 2.9 % higher in shallow), during the DST, the results are important as they indicate that crops in the shallow WTD areas have a delayed response to lower rainfall and have a relatively longer growing season. Further evidence of this delayed response can be seen in the EVI seasonality graphs (see Fig. S7.3), where the response of shallow crop to the DST seems delayed compared to deep areas. Crop LST further supports our hypothesis as LST in deep WTD areas

was  $0.39 \text{ }^\circ\text{C}$  higher than in shallow WTD areas, while no significant effects were found in EVI. Therefore cooler temperatures in shallow WTD areas are expected to be the result of higher evaporative cooling from ET. These relatively low differences in ET as measured with MODIS data might also be due to the ET product itself. The ET model used for MODIS is not optimised for comparison over relatively small spatial extents and short temporal scales (Ruhoff et al., 2013). In addition, the ET model does not take into account soil water storage and ET is based largely on atmospheric forcing and global land cover parameterisation. Therefore the differences we found for the DST may be underestimated in the MODIS ET values.

Ponte De Souza et al. (2011) highlighted that one of the strongest impacts of LULCC from forest to crop was due to the simultaneous 85 % increase in sensible heat flux and 78 % reduction in latent heat recorded during the dry season. Studies examining the change in LST due to LULCC found that LST increased by  $6 \text{ }^\circ\text{C}$  from forest to crop (Silvério et al., 2015) and  $1.5 \text{ }^\circ\text{C}$  from savanna to crop (Loarie et al., 2011). Further global models estimated an increase of  $5 \text{ }^\circ\text{C}$  during the summer season for the Amazon, due to a shift from forest to grass (Brovkin et al., 2009; Dekker et al., 2010). This increase in temperature could be influenced by WTD and land cover change; in shallow WTD areas this may result in a less severe temperature change, while in deep WTD it could lead to a greater change in temperature; however, WTD was not used as input for these modelling studies. Our results show a maximum temperature of  $30 \text{ }^\circ\text{C}$  in forest compared to a maximum temperature of  $38 \text{ }^\circ\text{C}$  in crops.

We also expected that the influence of WTD would be important during the WST, as in this period rainfall is increasing. In areas of shallow WTD, vegetation with a root zone close to the water table may still access water to supplement insufficient rainfall. Therefore, vegetation growth may be accelerated in comparison to areas of deep WTD which rely more directly on precipitation. Crop EVI was significantly higher in shallow than deep WTD areas by about 3.8 %, and these were the only data for which we found a significant difference. Looking at the seasonality of EVI (Fig. S7.3), during the WST EVI is increasing faster in shallow WTD areas than in deep WTD. EVI measures vegetation greenness and could be an indication of more rapid growth in shallow WTD areas. As EVI data are directly observed and not modelled, the differences are solely reliant on differences in reflected radiation. It may be that smaller differences between deep and shallow WTD areas are more easily detectable using these data. Along the arc of deforestation, observations of a lengthening dry season since the 1970s are linked to a delay in the WST (Butt et al., 2011; Fu et al., 2013). This delay correlates with LULCC and the large reduction this has on ET (DeBortoli et al., 2017). Although the difference in WTD seen in crops does not have a strong influence on ET when compared to the difference in ET between the land cover classes,

evidence of earlier or faster growth due to the shallow WTD could be beneficial on a local scale.

These results are even more relevant when comparing the effects of WTD in crop and forest. As forest has been shown to maintain ET throughout the seasons (Kunert et al., 2017) as its deep roots access deeper groundwater (Gash and Nobre, 1997; Nepstad et al., 1994), we hypothesised that no change should be observed in ET, LST, and EVI. Indeed, we found no significant differences across the three MODIS products, both annually or during the DST and WST. While this does not directly support our hypothesis about the role of WTD for shallow rooted vegetation, this does help validate that our approach reflects our knowledge of the system for vegetation with deep roots.

Savanna is a complex land cover type because its natural structure makes it challenging to classify with remote-sensing data (Gibbes et al., 2010). The MODIS classification accuracy of savanna is about 40 %, about half of that of forest and crop (90 % and 80 % respectively) (Friedl et al., 2010). Savanna includes both trees and grasses, which through the deep roots of trees may access moisture directly and facilitate moisture uptake via hydraulic redistribution (Oliveira et al., 2005), and large areas of shallow root grasses without trees would be negatively affected by water stress. A number of the findings for savanna were not in line with our proposed hypothesis. The distribution between shallow and deep LST was significantly different, with deep WTD areas having a skewed distribution towards lower temperatures. In our hypothesis, we expected to find lower temperature where shallow WTD occurs or no differences in temperature. A similar trend was found in ET where the 10th and 90th quantiles of the distribution were significantly higher in deep WTD areas. The difference in ET was very small, a less than 1 % difference between deep and shallow WTD areas. Water logging of soils has been shown to be an important factor in determining vegetation distribution (Ridolfi et al., 2006; Rossatto et al., 2012). Although we believe that larger flooding events leading to changes in vegetation composition are removed from our study due to the selection of pixels that during the time series were always classified as one land cover type, shorter periods of water logging may occur in shallow WTD areas. However, much higher spatial and temporal resolution imagery would be needed to identify this possibility.

The differences found for crop support our hypothesis that shallow WTD areas may facilitate water uptake compared with areas of deep WTD during the transition between wet and dry seasons. Previous crop production studies have shown that artificially maintaining a shallow WTD through subirrigation systems can increase the productivity of crops such as soy (Kahlow et al., 2005; Mejia et al., 2000), but this has not been previously shown in the naturally occurring shallow WTD areas of the arc of deforestation in the Amazon. In deep WTD areas, crop vegetation undergoes more severe water stress compared with shallow WTD, further reducing evapotranspiration and its potential impact on the

moisture recycling system. At the regional scale, the difference between deep and shallow WTD is not that important. The most significant differences in ET are driven by deforestation and strong annual variations in rainfall. Although not analysed specifically in this study, the remote-sensing data clearly show these distinctions between different land cover classes and high seasonal and inter-annual variability. On a local scale, the difference between deep and shallow WTD on crop may be of great importance. During the DST crop areas in shallow WTD maintained higher ET. This difference may be important for overall productivity as the dry season influence is delayed and, as a result, is increasing the growing season length. This could facilitate natural double cropping systems without the need for investment in irrigation, which is still an uncommon practice in the Amazon arc of deforestation (Lathuilière et al., 2012). Agricultural intensification is a pathway to increasing the sustainability of agriculture in the arc of deforestation if it prevents or reduces deforestation or facilitates reforestation (Oliveira et al., 2013). If agricultural productivity can be increased by focusing on already cleared shallow WTD areas, areas of deep WTD could be reforested or returned to secondary forest. Reforestation of previously degraded or logged forest has been shown to return to near-natural levels of ET within a few years (Davidson et al., 2012; Hölscher et al., 1997). The patterns seen in crop vegetation may be caused by factors not considered in this paper. Spatially explicit details about specific crops or agricultural practices were not known for the study. Planting of soybean is determined by the WST and can vary between September and October (Gusso et al., 2014). It is possible that the differences seen in shallow WTD could be the result of earlier sowing and double cropping systems. However, it may be that these agricultural management decisions are implemented more often in shallow WTD because of the higher availability of soil water.

This study is a first approach into gaining a better understanding of the influence of shallow WTD on shallow rooted vegetation, and it heavily relies on models and remote-sensing data which are most appropriate for analyses at larger spatial and temporal scales.

The results presented here are limited by the inherent uncertainty of the data used, both in the WTD model and in the remote-sensing data. Although we believe that the WTD model used here is the best currently available, due to limited data availability it was created using data located mostly in the coastal regions of the continent, with very few observations from near our study site (Fan and Miguez-Macho, 2010). In this study, the authors note that there is an overestimation of deep WTD areas when validated against literature-reported values. We believe that by the use of a conservative definition of deep WTD > 8 m the model outputs are appropriate for our purposes. As discussed above, the remote-sensing data have obvious limitations but do provide some insights into how depth of the water table at a local scale might affect water transfer and evaporative processes. Nonetheless,

the second main source of uncertainty is in the MODIS land cover classification. We chose to use this land cover classification as the ET and LST products use this classification in their algorithm. Although the classes used are broad and do not reflect the full complexity and heterogeneity of the arc of deforestation, they are robust enough for our purposes. As the influence of WTD on ET is most relevant on smaller scales, further research in these areas could focus on the smaller spatial scales and validate sites with accurate classification and WTD measurements.

## 5 Conclusion

This study aimed to investigate whether naturally occurring shallow water table depth supported higher ET compared to deep WTD, in particular, whether shallow rooted crop vegetation would have higher ET due to increased access to soil water in shallow WTD areas as the distance from the root zone to the saturated zone is shorter. Comparison of EVI showed evidence to support this hypothesis as daily mean EVI was significantly higher in shallow WTD crop areas. However, the difference between deep and shallow WTD is overshadowed by the clear differences between land cover classes. Although not the focus of this study, differences in ET, LST, and EVI were largest between land cover classes. In terms of larger-scale processes like moisture recycling, LULCC is far more impactful than WTD differences. The main driver of LULCC is agricultural expansion. So although our results are not directly relevant at regional or continental scales, on a local scale shallow WTD areas are more productive than deep WTD.

The influence of WTD on crop vegetation was concentrated during the transition periods between wet and dry seasons. We found higher ET and lower LST during the DST and higher EVI during the WST for crops in shallow WTD areas. This higher vegetation productivity of crops due to the shallow WTD helps to effectively increase the growing season length. The higher productivity in shallow WTD areas may facilitate natural double cropping, increasing the agricultural efficiency of the areas. These local-scale effects can become significant when scaled to the level of the Amazon. Deforestation rates grew as high as 28 000 km<sup>2</sup> yr<sup>-1</sup> in 2004 (Davidson et al., 2012). Any LULCC which occurs in areas of deep WTD leads to inefficiencies in agricultural production and higher impacts on the moisture recycling system.

The results presented here help to demonstrate that the LULCC impacts can vary spatially due to differences in WTD. Future studies investigating the impact of LULCC should incorporate WTD to help disentangle the full impact on the moisture recycling system.

*Data availability.* Data used in this paper is available at Fan and Miguez-Macho (2018) ([https://glowasis.deltares.nl/thredds/catalog/pendap/pendap/Equilibrium\\_Water\\_Table/catalog.html](https://glowasis.deltares.nl/thredds/catalog/pendap/pendap/Equilibrium_Water_Table/catalog.html)).

*Supplement.* The supplement related to this article is available online at: <https://doi.org/10.5194/hess-23-3917-2019-supplement>.

*Author contributions.* All the authors contributed to the study in design and implementation. The original draft was written by JOC, and all the authors contributed with revisions and editing.

*Competing interests.* The authors declare that they have no conflict of interest.

*Acknowledgements.* This study was supported by funding from the “Nature Conservation, Management and Restoration” graduate programme of the Netherlands Organisation for Scientific Research (NWO). We would like to thank the reviewers for their helpful comments throughout the discussion process.

*Financial support.* This research has been supported by the Nederlandse Organisatie voor Wetenschappelijk Onderzoek (grant no. 022.005.028).

*Review statement.* This paper was edited by Ryan Teuling and reviewed by two anonymous referees.

## References

- Arvor, D., Dubreuil, V., Ronchail, J., Simões, M., and Funatsu, B. M.: Spatial patterns of rainfall regimes related to levels of double cropping agriculture systems in Mato Grosso (Brazil), *Int. J. Climatol.*, 34, 2622–2633, <https://doi.org/10.1002/joc.3863>, 2014.
- Asner, G. P., Scurlock, J. M. O., and Hicke, A. J.: Global synthesis of leaf area index observations: implications for ecological and remote sensing studies, *Global Ecol. Biogeogr.*, 12, 191–205, <https://doi.org/10.1046/j.1466-822X.2003.00026.x>, 2003.
- Badger, A. M. and Dirmeyer, P. A.: Climate response to Amazon forest replacement by heterogeneous crop cover, *Hydrol. Earth Syst. Sci.*, 19, 4547–4557, <https://doi.org/10.5194/hess-19-4547-2015>, 2015.
- Bonan, G. B.: Forests and climate change: forcings, feedbacks, and the climate benefits of forests, *Science*, 320, 1444–1449, <https://doi.org/10.1126/science.1155121>, 2008.
- Brando, P. M., Balch, J. K., Nepstad, D. C., Morton, D. C., Putz, F. E., Coe, M. T., Silverio, D., Macedo, M. N., Davidson, E. A., Nobrega, C. C., Alencar, A., and Soares-Filho, B. S.: Abrupt increases in Amazonian tree mortality due to drought-fire interactions, *P. Natl. Acad. Sci. USA*, 111, 6347–6352, <https://doi.org/10.1073/pnas.1305499111>, 2014.

- Brovkin, V., Raddatz, T., Reick, C. H., Claussen, M., and Gayler, V.: Global biogeophysical interactions between forest and climate, *Geophys. Res. Lett.*, 36, 1–5, <https://doi.org/10.1029/2009GL037543>, 2009.
- Butt, N., de Oliveira, P. A., and Costa, M. H.: Evidence that deforestation affects the onset of the rainy season in Rondonia, Brazil, *J. Geophys. Res.*, 116, D11120, <https://doi.org/10.1029/2010JD015174>, 2011.
- Canadell, J., Jackson, R., Ehleringer, J., Mooney, H. A., Sala, O. E., and Schulze, E.-D.: Maximum rooting depth of vegetation types at the global scale, *Oecologia*, 108, 583–595, <https://doi.org/10.1007/BF00329030>, 1996.
- Chambers, J. Q., Asner, G. P., Morton, D. C., Anderson, L. O., Saatchi, S. S., Espírito-Santo, F. D. B., Palace, M., and Souza, C.: Regional ecosystem structure and function: ecological insights from remote sensing of tropical forests, *Trends Ecol. Evol.*, 22, 414–423, <https://doi.org/10.1016/j.tree.2007.05.001>, 2007.
- Costa, M. H. and Foley, J. A.: Combined effects of deforestation and doubled atmospheric CO<sub>2</sub> concentrations on the climate of Amazonia, *J. Climate*, 13, 18–34, [https://doi.org/10.1175/1520-0442\(2000\)013<0018:CEODAD>2.0.CO;2](https://doi.org/10.1175/1520-0442(2000)013<0018:CEODAD>2.0.CO;2), 2000.
- Costa, M. H. and Pires, G. F.: Effects of Amazon and Central Brazil deforestation scenarios on the duration of the dry season in the arc of deforestation, *Int. J. Climatol.*, 30, 1970–1979, <https://doi.org/10.1002/joc.2048>, 2010.
- Costa, M. H., Yanagi, S. N. M., Souza, P. J. O. P., Ribeiro, A., and Rocha, E. J. P.: Climate change in Amazonia caused by soybean cropland expansion, as compared to caused by pastureland expansion, *Geophys. Res. Lett.*, 34, 2–5, <https://doi.org/10.1029/2007GL029271>, 2007.
- Davidson, E. A., de Araújo, A. C., Artaxo, P., Balch, J. K., Brown, I. F. C., Bustamante, M. M., Coe, M. T., DeFries, R. S., Keller, M., Longo, M., Munger, J. W., Schroeder, W., Soares-Filho, B. S., Souza, C. M., and Wofsy, S. C.: The Amazon basin in transition, *Nature*, 481, 321–328, <https://doi.org/10.1038/nature10717>, 2012.
- Debortoli, N. S., Dubreuil, V., Hirota, M., Filho, S. R., Lindoso, D. P., and Nabucet, J.: Detecting deforestation impacts in Southern Amazonia rainfall using rain gauges, *Int. J. Climatol.*, 37, 2889–2900, <https://doi.org/10.1002/joc.4886>, 2017.
- Dekker, S. C., de Boer, H. J., Brovkin, V., Fraedrich, K., Wassen, M. J., and Rietkerk, M.: Biogeophysical feedbacks trigger shifts in the modelled vegetation–atmosphere system at multiple scales, *Biogeosciences*, 7, 1237–1245, <https://doi.org/10.5194/bg-7-1237-2010>, 2010.
- Dubreuil, V., Debortoli, N., Funatsu, B., Nédélec, V., and Durieux, L.: Impact of land-cover change in the Southern Amazonia climate: a case study for the region of Alta Floresta, Mato Grosso, Brazil, *Environ. Monit. Assess.*, 184, 877–891, <https://doi.org/10.1007/s10661-011-2006-x>, 2012.
- Eltahir, E. A. B. and Bras, R. L.: Precipitation recycling in the Amazon basin, *Q. J. Roy. Meteorol. Soc.*, 120, 861–880, <https://doi.org/10.1002/qj.49712051806>, 1994a.
- Eltahir, E. A. B. and Bras, R. L.: Sensitivity of regional climate to deforestation in the Amazon basin, *Adv. Water Resour.*, 17, 101–115, [https://doi.org/10.1016/0309-1708\(94\)90027-2](https://doi.org/10.1016/0309-1708(94)90027-2), 1994b.
- Fan, J., McConkey, B., Wang, H., and Janzen, H.: Root distribution by depth for temperate agricultural crops, *F. Crop. Res.*, 189, 68–74, <https://doi.org/10.1016/j.fcr.2016.02.013>, 2016.
- Fan, Y. and Miguez-Macho, G.: Potential groundwater contribution to Amazon evapotranspiration, *Hydrol. Earth Syst. Sci.*, 14, 2039–2056, <https://doi.org/10.5194/hess-14-2039-2010>, 2010.
- Fan, Y. and Miguez-Macho, G.: Equilibrium water table depth for South America, Earthdata, available at: [https://glowasis.deltares.nl/thredds/catalog/opendap/opendap/Equilibrium\\_Water\\_Table/catalog.html](https://glowasis.deltares.nl/thredds/catalog/opendap/opendap/Equilibrium_Water_Table/catalog.html), last access: April 2018.
- Fearnside, P. M.: Environmental services as a strategy for sustainable development in rural Amazonia, *Ecol. Econ.*, 20, 53–70, [https://doi.org/10.1016/S0921-8009\(96\)00066-3](https://doi.org/10.1016/S0921-8009(96)00066-3), 1997.
- Foley, J. A., Asner, G. P., Costa, M. H., Coe, M. T., DeFries, R., Gibbs, H. K., Howard, E. A., Olson, S., Patz, J., Ramankutty, N., and Snyder, P.: Amazonia revealed: forest degradation and loss of ecosystem goods and services in the Amazon Basin, *Front. Ecol. Environ.*, 5, 25–32, [https://doi.org/10.1890/1540-9295\(2007\)5\[25:ARFDAL\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2007)5[25:ARFDAL]2.0.CO;2), 2007.
- Friedl, M. A., Sulla-Menashe, D., Tan, B., Schneider, A., Ramankutty, N., Sibley, A., and Huang, X.: MODIS Collection 5 global land cover: Algorithm refinements and characterization of new datasets, *Remote Sens. Environ.*, 114, 168–182, <https://doi.org/10.1016/j.rse.2009.08.016>, 2010.
- Fu, R., Yin, L., Li, W., Arias, P. A., Dickinson, R. E., Huang, L., Chakraborty, S., Fernandes, K., Liebmann, B., Fisher, R., and Myneni, R. B.: Increased dry-season length over southern Amazonia in recent decades and its implication for future climate projection, *P. Natl. Acad. Sci. USA*, 110, 18110–18115, <https://doi.org/10.1073/pnas.1302584110>, 2013.
- Gash, J. H. C. and Nobre, C. A.: Climatic Effects of Amazonian Deforestation: Some Results from ABRACOS, *B. Am. Meteorol. Soc.*, 78, 823–830, [https://doi.org/10.1175/1520-0477\(1997\)078<0823:CEOADS>2.0.CO;2](https://doi.org/10.1175/1520-0477(1997)078<0823:CEOADS>2.0.CO;2), 1997.
- Gibbes, C., Adhikari, S., Rostant, L., Southworth, J., and Qiu, Y.: Application of Object Based Classification and High Resolution Satellite Imagery for Savanna Ecosystem Analysis, *Remote Sens.*, 2, 2748–2772, <https://doi.org/10.3390/rs2122748>, 2010.
- Gusso, A., Arvor, D., Ricardo Ducati, J., Veronez, M. R., and Da Silveira, L. G.: Assessing the modis crop detection algorithm for soybean crop area mapping and expansion in the Mato Grosso state, Brazil, *Sci. World J.*, 2014, 863141, <https://doi.org/10.1155/2014/863141>, 2014.
- Hölscher, D., de Sá, T. A., Bastos, T., Denich, M., and Fölster, H.: Evaporation from young secondary vegetation in eastern Amazonia, *J. Hydrol.*, 193, 293–305, [https://doi.org/10.1016/S0022-1694\(96\)03145-9](https://doi.org/10.1016/S0022-1694(96)03145-9), 1997.
- Huete, A., Didan, K., Miura, H., Rodriguez, E. P., Gao, X., and Ferreira, L. F.: Overview of the radiometric and biophysical performance of the MODIS vegetation indices, *Remote Sens. Environ.*, 83, 195–213, 2002.
- Kahlow, M. A., Ashraf, M., and Zia-Ul-Haq: Effect of shallow groundwater table on crop water requirements and crop yields, *Agr. Water Manage.*, 76, 24–35, <https://doi.org/10.1016/j.agwat.2005.01.005>, 2005.
- Kastens, J. H., Brown, J. C., Coutinho, A. C., Bishop, C. R., and Esquerdo, J. C. D. M.: Soy moratorium impacts on soybean and deforestation dynamics in Mato Grosso, Brazil, *PLoS One*, 12, 1–21, <https://doi.org/10.1371/journal.pone.0176168>, 2017.
- Kunert, N., Maria Aparecido, L. T., Wolff, S., Higuchi, N., Dos, J., Carioca de Araujo, A., and Trumbore, S.: A revised hydrological model for the Central Amazon: The importance of emergent

- canopy trees in the forest water budget, *Agr. Forest Meteorol.*, 239, 47–57, <https://doi.org/10.1016/j.agrformet.2017.03.002>, 2017.
- Lathuilière, M. J., Johnson, M. S., and Donner, S. D.: Water use by terrestrial ecosystems: temporal variability in rainforest and agricultural contributions to evapotranspiration in Mato Grosso, Brazil, *Environ. Res. Lett.*, 7, 024024, <https://doi.org/10.1088/1748-9326/7/2/024024>, 2012.
- Lawrence, D. and Vandecar, K.: Effects of tropical deforestation on climate and agriculture, *Nat. Clim. Change*, 5, 27–36, <https://doi.org/10.1038/nclimate2430>, 2014.
- Liebmann, B., Camargo, S. J., Seth, A., Marengo, J. A., Carvalho, L. M. V., Allured, D., Fu, R., and Vera, C. S.: Onset and end of the rainy season in South America in observations and the ECHAM 4.5 atmospheric general circulation model, *J. Climate*, 20, 2037–2050, <https://doi.org/10.1175/JCLI4122.1>, 2007.
- Loarie, S. R., Lobell, D. B., Asner, G. P., Mu, Q., and Field, C. B.: Direct impacts on local climate of sugarcane expansion in Brazil, *Nat. Clim. Change*, 1, 105–109, <https://doi.org/10.1038/nclimate1067>, 2011.
- Maeda, E. E., Ma, X., Wagner, F. H., Kim, H., Oki, T., Eamus, D., and Huete, A.: Evapotranspiration seasonality across the Amazon Basin, *Earth Syst. Dynam.*, 8, 439–454, <https://doi.org/10.5194/esd-8-439-2017>, 2017.
- Makarieva, A. M. and Gorshkov, V. G.: Biotic pump of atmospheric moisture as driver of the hydrological cycle on land, *Hydrol. Earth Syst. Sci.*, 11, 1013–1033, <https://doi.org/10.5194/hess-11-1013-2007>, 2007.
- Malhi, Y., Roberts, J. T., Betts, R. A., Killeen, T. J., Li, W., and Nobre, C. A.: Climate Change, Deforestation, and the Fate of the Amazon, *Science*, 319, 169–172, <https://doi.org/10.1126/science.1146961>, 2008.
- Mejia, M. N., Madramootoo, C. A., and Broughton, R. S.: Influence of water table management on corn and soybean yields, *Agr. Water Manage.*, 46, 73–89, [https://doi.org/10.1016/S0378-3774\(99\)00109-2](https://doi.org/10.1016/S0378-3774(99)00109-2), 2000.
- Miguez-Macho, G. and Fan, Y.: The role of groundwater in the Amazon water cycle: 1. Influence on seasonal streamflow, flooding and wetlands, *J. Geophys. Res.-Atmos.*, 117, 1–30, <https://doi.org/10.1029/2012JD017539>, 2012.
- Moreira, M. Z., Sternberg, L. D. L., and Nepstad, D. C.: Vertical patterns of soil water uptake by plants in a primary forest and an abandoned pasture in the eastern Amazon: an isotopic approach, *Plant Soil*, 222, 95–107, <https://doi.org/10.1023/A:1004773217189>, 2000.
- Mu, Q., Zhao, M., and Running, S. W.: Improvements to a MODIS global terrestrial evapotranspiration algorithm, *Remote Sens. Environ.*, 115, 1781–1800, <https://doi.org/10.1016/j.rse.2011.02.019>, 2011.
- Neill, C., Coe, M. T., Riskin, S. H., Krusche, A. V., Eisenbeer, H., Macedo, M. N., McHorney, R., Lefebvre, P., Davidson, E. A., Scheffler, R., Figueira, A. M. A. S., Porder, S., and Deegan, L. A.: Watershed responses to Amazon soya bean cropland expansion and intensification, *Philos. T. Roy. Soc. Lond. B*, 368, 2–7, <https://doi.org/10.1098/rstb.2012.0425>, 2013.
- Nepstad, D. C., de Carvalho, C. R., Davidson, E. A., Jipp, P. H., Lefebvre, P. A., Negreiros, G. H., da Silva, E. D., Stone, T. A., Trumbore, S. E., and Vieira, S.: The role of deep roots in the hydrological and carbon cycles of Amazonian forests and pastures, *Nature*, 372, 666–669, <https://doi.org/10.1038/372666a0>, 1994.
- Nobre, C. A., Sampaio, G., Borma, L. S., Castilla-Rubio, J. C., Silva, J. S., and Cardoso, M.: Land-use and climate change risks in the Amazon and the need of a novel sustainable development paradigm, *P. Natl. Acad. Sci. USA*, 113, 10759–10768, <https://doi.org/10.1073/pnas.1605516113>, 2016a.
- Nobre, C. A., Marengo, J. A., Seluchi, M. E., Cuartas, A., and Alves, L. M.: Some Characteristics and Impacts of the Drought and Water Crisis in Southeastern Brazil during 2014 and 2015 Some Characteristics and Impacts of the Drought and Water Crisis in Southeastern Brazil during 2014 and 2015, *J. Water Resour. Prot.*, 8, 252–262, <https://doi.org/10.4236/jwarp.2016.82022>, 2016b.
- Oliveira, L. J. C., Costa, M. H., Soares-Filho, B. S., and Coe, M. T.: Large-scale expansion of agriculture in Amazonia may be a no-win scenario, *Environ. Res. Lett.*, 8, 024021, <https://doi.org/10.1088/1748-9326/8/2/024021>, 2013.
- Oliveira, R. S., Dawson, T. E., Burgess, S. S. O., and Nepstad, D. C.: Hydraulic redistribution in three Amazonian trees, *Oecologia*, 145, 354–363, <https://doi.org/10.1007/s00442-005-0108-2>, 2005.
- Ponte De Souza, P. J. D. O., Ribeiro, A., Da Rocha, E. J. P., Botelho, M. D. N., De Sousa, A. M. L., De Souza, E. B., and Bouças Farias, J. R.: Impacts of soybean expansion on the Amazon energy balance: A case study, *Exp. Agric.*, 47, 553–567, <https://doi.org/10.1017/S0014479711000391>, 2011.
- Ridolfi, L., Odorico, P. D., and Laio, F.: Effect of vegetation–water table feedbacks on the stability and resilience of plant ecosystems, *Water Resour. Res.*, 42, 1–5, <https://doi.org/10.1029/2005WR004444>, 2006.
- Rossatto, D. R., de Carvalho Ramos Silva, L., Villalobos-Vega, R., Sternberg, L., Da, S. L., and Franco, A. C.: Depth of water uptake in woody plants relates to groundwater level and vegetation structure along a topographic gradient in a neotropical savanna, *Environ. Exp. Bot.*, 77, 259–266, <https://doi.org/10.1016/j.envexpbot.2011.11.025>, 2012.
- Ruhoff, A. L., Paz, A. R., Aragao, L. E. O. C., Mu, Q., Malhi, Y., Collischonn, W., Rocha, H. R., and Running, S. W.: Assessment of the MODIS global evapotranspiration algorithm using eddy covariance measurements and hydrological modelling in the Rio Grande basin, *Hydrolog. Sci. J.*, 58, 1658–1676, <https://doi.org/10.1080/02626667.2013.837578>, 2013.
- Sampaio, G., Nobre, C., Costa, M. H., Satyamurty, P., Soares-Filho, B. S., and Cardoso, M.: Regional climate change over eastern Amazonia caused by pasture and soybean cropland expansion, *Geophys. Res. Lett.*, 34, L17709, <https://doi.org/10.1029/2007GL030612>, 2007.
- Setiyono, T. D., Weiss, A., Specht, J. E., Cassman, K. G., and Dobermann, A.: Leaf area index simulation in soybean grown under near-optimal conditions, *F. Crop. Res.*, 108, 82–92, <https://doi.org/10.1016/j.fcr.2008.03.005>, 2008.
- Sheil, D.: How plants water our planet: advances and imperatives, *Trends Plant Sci.*, 19, 209–211, <https://doi.org/10.1016/j.tplants.2014.01.002>, 2014.
- Silvério, D. V., Brando, P. M., Macedo, M. N., Beck, P. S. A., Bustamante, M., and Coe, M. T.: Agricultural expansion dominates climate changes in southeastern Amazonia: The over-

- looked non-GHG forcing, *Environ. Res. Lett.*, 10, 104015, <https://doi.org/10.1088/1748-9326/10/10/104015>, 2015.
- Spracklen, D. V., Arnold, S. R., and Taylor, C. M.: Observations of increased tropical rainfall preceded by air passage over forests, *Nature*, 489, 282–285, <https://doi.org/10.1038/nature11390>, 2012.
- Staal, A., Tuinenburg, O. A., Bosmans, J. H. C., Holmgren, M., van Nes, E. H., Scheffer, M., Zemp, D. C., and Dekker, S. C.: Forest-rainfall cascades buffer against drought across the Amazon, *Nat. Clim. Change*, 8, 539–543, <https://doi.org/10.1038/s41558-018-0177-y>, 2018.
- Swann, A. L. S., Longo, M., Knox, R. G., Lee, E., and Moorcroft, P. R.: Future deforestation in the Amazon and consequences for South American climate, *Agr. Forest Meteorol.*, 214–215, 12–24, <https://doi.org/10.1016/j.agrformet.2015.07.006>, 2015.
- van der Ent, R. J. and Savenije, H. H. G.: Length and time scales of atmospheric moisture recycling, *Atmos. Chem. Phys.*, 11, 1853–1863, <https://doi.org/10.5194/acp-11-1853-2011>, 2011.
- Velpuri, N. M., Senay, G. B., Singh, R. K., Bohms, S., and Verdin, J. P.: A comprehensive evaluation of two MODIS evapotranspiration products over the conterminous United States: Using point and gridded FLUXNET and water balance ET, *Remote Sens. Environ.*, 139, 35–49, <https://doi.org/10.1016/j.rse.2013.07.013>, 2013.
- Vergopolan, N. and Fisher, J. B.: The impact of deforestation on the hydrological cycle in Amazonia as observed from remote sensing, *Int. J. Remote Sens.*, 37, 5412–5430, <https://doi.org/10.1080/01431161.2016.1232874>, 2016.
- Wan, Z.: New refinements and validation of the collection-6 MODIS land-surface temperature/emissivity product, *Remote Sens. Environ.*, 140, 36–45, <https://doi.org/10.1016/j.rse.2013.08.027>, 2014.
- Wilcox, R. R.: Comparing two independent groups via a quantile generalization of the wilcoxon-mann-whitney test, *J. Mod. Appl. Stat. Meth.*, 11, 296–302, <https://doi.org/10.22237/jmasm/1351742460>, 2012.
- Wright, J. S., Fu, R., Worden, J. R., Chakraborty, S., Clinton, N. E., Risi, C., Sun, Y., and Yin, L.: Rainforest-initiated wet season onset over the southern Amazon, *P. Natl. Acad. Sci. USA*, 114, 8481–8486, <https://doi.org/10.1073/pnas.1621516114>, 2017.
- Xu, D., Agee, E., Wang, J., and Ivanov, V. Y.: Estimation of Evapotranspiration of Amazon Rainforest Using the Maximum Entropy Production Method, *Geophys. Res. Lett.*, 1402–1412, <https://doi.org/10.1029/2018gl080907>, 2019.
- Zalles, V., Hansen, M. C., Potapov, P. V., Stehman, S. V., Tyukavina, A., Pickens, A., Song, X., Adusei, B., Okpa, C., Aguilar, R., John, N., and Chavez, S.: Near doubling of Brazil's intensive row crop area since 2000 [Sustainability Science], *P. Natl. Acad. Sci. USA*, 116, 428–435, <https://doi.org/10.1073/pnas.1810301115>, 2019.